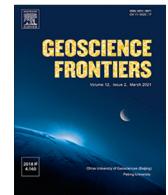




Contents lists available at ScienceDirect



## Research Paper

# Physiological characteristics, geochemical properties and hydrological variables influencing pathogen migration in subsurface system: What we know or not?

Wenjing Zhang <sup>a,b</sup>, Juanfen Chai <sup>a,b</sup>, Shuxin Li <sup>a,b</sup>, Xinzi Wang <sup>c</sup>, Shengyu Wu <sup>a,b</sup>, Zhentian Liang <sup>a,b</sup>, Muhammad Yousuf Jat Baloch <sup>a,b</sup>, Luis F.O. Silva <sup>d,e</sup>, Dayi Zhang <sup>a,b,c,\*</sup>

<sup>a</sup> Key Laboratory of Groundwater Resources and Environment (Jilin University), Ministry of Education, Changchun 130021, China

<sup>b</sup> College of New Energy and Environment, Jilin University, Changchun 130021, China

<sup>c</sup> School of Environment, Tsinghua University, Beijing 100084, China

<sup>d</sup> Departamento de Ingeniería Civil y Arquitectura, Universidad de Lima, Avenida Javier Prado Este 4600, Santiago de Surco 1503, Peru

<sup>e</sup> Department of Civil and Environmental, Universidad de la Costa, Calle 58 #55-66, 080002 Barranquilla, Atlántico, Colombia

## ARTICLE INFO

## Article history:

Received 2 September 2021

Revised 23 October 2021

Accepted 30 December 2021

Available online xxxx

## Keywords:

Pathogens

Viruses

SARS-CoV-2

Subsurface system

Migration

## ABSTRACT

The global outbreak of coronavirus infectious disease-2019 (COVID-19) draws attentions in the transport and spread of Severe Acute Respiratory Syndrome Coronavirus 2 (SARS-CoV-2) in aerosols, wastewater, surface water and solid wastes. As pathogens eventually enter the subsurface system, e.g., soils in the vadose zone and groundwater in the aquifers, they might survive for a prolonged period of time owing to the uniqueness of subsurface environment. In addition, pathogens can transport in groundwater and contaminate surrounding drinking water sources, possessing long-term and concealed risks to human society. This work critically reviews the influential factors of pathogen migration, unravelling the impacts of pathogenic characteristics, vadose zone physiochemical properties and hydrological variables on the migration of typical pathogens in subsurface system. An assessment algorithm and two rating/weighting schemes are proposed to evaluate the migration abilities and risks of pathogens in subsurface environment. As there is still no evidence about the presence and distribution of SARS-CoV-2 in the vadose zones and aquifers, this study also discusses the migration potential and behavior of SARS-CoV-2 viruses in subsurface environment, offering prospective clues and suggestions for its potential risks in drinking water and effective prevention and control from hydrogeological points of view.

© 2021 China University of Geosciences (Beijing) and Peking University. Production and hosting by Elsevier B.V. This is an open access article under the CC BY-NC-ND license (<http://creativecommons.org/licenses/by-nc-nd/4.0/>).

## 1. Introduction

Since the beginning of the 21st century, there are increasing emerging diseases caused by various pathogens, including swine-origin pandemic 2009 A influenza (H1N1), Ebola and Nipah (Murphy, 1998; Morse et al., 2012). The recent outbreak of coronavirus infectious disease 2019 (COVID-19) is caused by severe acute respiratory syndrome coronavirus 2 (SARS-CoV-2) and becomes a public health emergency of international concerns causing over 240 million cases and nearly 5 million deaths by 18th October 2021 in 195 countries. As most pathogens can be released by patients into the surrounding environment and survive for a prolonged period of time, increasing concerns have been raised

for their potential transport and spread via environmental media. For instance, noroviruses causing acute gastroenteritis mainly infect humans by contaminated foods (Cheng et al., 2005; Widdowson et al., 2005) and are frequently detected in water environment after an outbreak (Haramoto et al., 2005; Rutjes et al., 2005). SARS-CoV-2 transmits mainly through aerosols, fomites (Sosnowski, 2021) and respiratory droplets (Kampf et al., 2020). Additionally, there is increasing evidence showing the potential spillover of pathogens from natural hosts in their wildlife habitats to human societies across the ecological barriers (Zhang et al., 2021a), or from dedicated hospitals to surrounding public communities via aerosols (Zhang et al., 2021b) or wastewater (Randazzo et al., 2020; Zhang et al., 2020). Thus, it is of urgency to study the migration of pathogens, especially those causing severe pandemics like Nipah and SARS-CoV-2, in natural environment for effective prevention and control of emerging infectious diseases.

\* Corresponding author at: Key Laboratory of Groundwater Resources and Environment (Jilin University), Ministry of Education, Changchun 130021, China.

E-mail address: [zhangdayi@tsinghua.org.cn](mailto:zhangdayi@tsinghua.org.cn) (D. Zhang).

Groundwater is widely used for drinking water supply around the world because of freshwater scarcity (Murphy et al., 2017; Panda et al., 2020). Approximately 2.2 billion people globally rely on groundwater for daily use (Murphy et al., 2017), and over 70% of drinking water supplies in California's San Joaquin Valley are groundwater (Dieter and Maupin, 2017). Due to the intensive irrigation and improper sewage treatment, pathogens can migrate into subsurface environment and are frequently detected in aquifers, causing serious pollution and imposing risks for drinking water safety (Flury and Aramrak, 2017; Haveman, 2002; Makowska et al., 2021). Contaminated groundwater will be ingested as drinking water, during recreation or via irrigated crops (Oliveira et al., 2016). Pathogens in groundwater are estimated to cause 35.2–59.4 million cases of acute gastrointestinal illness and over 2 million deaths worldwide annually (Pandey et al., 2014). Waterborne enteric pathogens are responsible for diarrhea which is the second leading diseases causing infant deaths worldwide (Wardlaw et al., 2010). Over 21 thousand cases of water-borne diseases occurred in USA during 1999–2019, including vector-borne diseases and water-borne toxins (Rhoden et al., 2021), and contaminated groundwater accounts for about 64% of the drinking water outbreaks between 1989 and 2002 (Fong et al., 2007). The norovirus outbreak in Wisconsin in June 2007 is caused by fecal pathogens in contaminated aquifers (Borchardt et al., 2011). In Delhi, about 28% of the household were affected by water-borne diseases (Bidhuri et al., 2018). In Norway, the proportion of water-borne outbreaks related to groundwater reflects the proportion of groundwater works from 1984 to 2007, and the proportion of disease cases caused by groundwater corresponds to the proportion of persons supplied by groundwater (Kvitsand and Fiksdal, 2010). Therefore, pathogens in groundwater are key sources of infectious diseases and threaten the public health, and it is necessary to study the existence and migration of pathogens in subsurface environment.

Extensive studies have reported the occurrence of pathogens in groundwater, including bacteria and viruses (Hynds et al., 2014). In Jianghan Plain of China, 193 potentially pathogenic bacterial species are identified in groundwater, and opportunistic pathogenic *Acinetobacter baumannii*-like and enteric pathogenic *Vibrio cholerae*-like bacteria are present in 99% and 32% of groundwater samples, respectively (Wu et al., 2019). Frequently detected pathogenic viruses in groundwater include enterovirus, hepatitis virus, norovirus, adenovirus and rotavirus, and they are identified in 20%–30% of 550 wells in the United States and urban aquifers in the United Kingdom (Foster and Chilton, 2004). From 1992 to 2013, one or more viruses are detected in 15% of 2273 groundwater samples from 746 public drinking water systems in Canada (Fout et al., 2017), and human enterovirus (polyomavirus, adenovirus A, and GII norovirus) are detected in 10 of 22 wells (Allen et al., 2017). In Korea, 35.3%–48.3% of 71 groundwater sites are positive for waterborne norovirus (Lee et al., 2011). Particularly, the living pathogenic viruses are even detected in aquifers up to 40 m at depth in the UK (Cronin et al., 2003).

Besides the distribution and composition of pathogens in groundwater, many studies also focus on their migration and transport in subsurface environment. Pathogen migration from above-ground is believed as the main source of groundwater pathogenic contamination (Alegbeleye and Sant'Ana, 2020), and the vertical transport is mainly accompanied with the rainfall and wastewater infiltration by agricultural irrigation and improper sewage discharge (Bellou et al., 2015; Syngouna et al., 2017; Alegbeleye and Sant'Ana, 2020). The abundance of *Escherichia coli* in groundwater increases by one to three orders of magnitude after rainfall in karst terrain (Buckerfield et al., 2019, 2020). The agricultural irrigation and pit latrine are the main sources of underground pathogens in developing countries (Liu et al., 2021a), and norovirus

is detected in 1% (7 of 773) of irrigation waters in South Korea (Shin et al., 2019). In addition, pathogens are abundant in feces or urine of humans or animals (Hellmér et al., 2014; Gerba et al., 2017; Ong et al., 2020; Sun et al., 2020), and the urban sewage or animal wastes are the main routes of pathogenic transmission into the environment (Fumian et al., 2010). Globally, about 100 billion tons of animal wastes are produced annually as soil conditioners or crop fertilizers (Pennington, 2010), and many pathogens might be released into agricultural fields. During the COVID-19, studies have confirmed the presence of SARS-CoV-2 in untreated wastewater (Ahmed et al., 2020; La Rosa et al., 2020; Randazzo et al., 2020; Sherchan et al., 2020; Zhang et al., 2021c). Considering the potential transmission from the urban sewage to humans and animals via natural environment (Rizzo et al., 2013; Fouz et al., 2020) and leakage into the underground environment directly in rural area (Rashid and Pandit, 2017; Weststrate et al., 2019; Buckerfield et al., 2020; Li et al., 2021), it is important to study the mechanisms of pathogen migration and transport in the vadose zone.

Pathogenic transport in subsurface environment depends on many geochemical variables (e.g., grain size, porosity, heterogeneity, moisture content, pH, ionic strength, temperature) and hydrodynamics (e.g., flow rate and direction) (Jin and Flury, 2002; Bradford et al., 2013; Qin et al., 2020; Zhang et al., 2021e). The activities of some pathogens decreases with temperature (Balboa et al., 2021; Gundy et al., 2009; Pinon and Viallette, 2018), and pathogen migration is benefited by a higher pH value (Torkzaban et al., 2006). In the vadose zone, the migration of pathogens is closely related to the grain size and moisture content (Anders and Chrysikopoulos, 2006; Bai et al., 2016), whereas in heterogeneous aquifer media, it depends mainly on the flow rate and flow direction (Luo et al., 2020). In the presence of preferential migration pathway in vadose zone or aquifer, pathogens can even migrate faster (Corapcioglu et al., 2006). In addition, pathogen migration is also strongly associated with pathogenic physiological characteristics like size, morphology, and hydrophobicity (Ghanem et al., 2016; Pang et al., 2021). Pathogens with smaller size, higher hydrophilic and spherical shapes are more favored for migration (Weiss et al., 1995; Schinner et al., 2010; Whitehead and McCue, 2010; Farkas et al., 2015; Schijven and Hassanizadeh, 2000; Pang et al., 2021). Nevertheless, most studies only focus on and compare limited pathogenic species or influential factors, and there is still lack of a comprehensive and complete understanding on the influential factors and risks of pathogen migration in subsurface environment.

This work critically reviews the influential factors of pathogen migration in subsurface environment, attempting to discuss the impacts of pathogenic physiological characteristics and geochemical/hydrological variables on the transport and spread of typical pathogens in the vadose and saturated zones. More importantly, we propose a database of pathogenic physiological characteristics and an assessment algorithm to evaluate the migration ability and risks of pathogens in subsurface environment. Our findings can predict the potential behavior of SARS-CoV-2 in subsurface environment, provide prospective clues for its potential risks in groundwater and drinking water, and offer suggestions for effective prevention and control of emerging infectious diseases from hydrogeological points of view.

## 2. Pathogen physiological characteristics

With the recent development of virology and genetics, the unique morphological characteristics and genetic diversity of many pathogens have been widely uncovered and linked to their behavior in environment (Dion et al., 2020; Zimmerman et al.,

2020). Generally, pathogen size, morphological features, surface charge, surface hydrophobicity and membrane structure are the most investigated and accepted physiological characteristics affecting pathogen migration in subsurface environment.

### 2.1. Pathogen size

Pathogen size is a critical factor for their transport in porous media. As particles with smaller size pass filters such as soils and sediments more easily, pathogenic biocolloids with larger size migrate slower by colliding with and adhering to porous medium (Schijven and Hassanzadeh, 2000; Pelley and Tufenkji, 2008; Aronino et al., 2009; Schinner et al., 2010). For bacteria which in general have larger sizes (0.5–3  $\mu\text{m}$ ) than viruses (20–90 nm) (Walsh et al., 2010), they are more readily to be retained in porous media by straining. Viruses therefore normally migrate faster in porous media, generally two to three times than pathogenic bacteria (Robertson and Edberg, 1997). Sinton et al. found a significantly different transport and attenuation behavior between bacterial *Escherichia coli* (1.0–1.5  $\mu\text{m}$  in diameter and 3.0–5.0  $\mu\text{m}$  in length) and viral F-RNA bacteriophage (26 nm in diameter, the same as follows) (Sinton et al., 2010). The transport of viruses MS-2 and PRD-1 in porous media had a faster peak emergence than bacteria *Salmonella* and *Klebsiella*, although *Salmonella* and *Klebsiella* survive for a longer time (dropped by 1–6 log units in 32 d, whereas viruses below the detection limit within 10 d) (Dowd and Pillai, 1997). For bacterial pathogens of similar size like *Escherichia coli* (1.11  $\mu\text{m}$ ), *Klebsiella* sp. (1.56  $\mu\text{m}$ ) and *Rhodococcus rhodochrous* (2.31  $\mu\text{m}$ ), their transport is more dependent on grain size (Bai et al., 2016). Although smaller colloids with size <1  $\mu\text{m}$  may be easier to deposit on medium surfaces for their greater Brownian diffusion, most studies document that viruses with smaller size migrate faster. Particularly, 60 nm is reported as the critical diameter of viral transportation, below which viral migration is in a considerably longer distance (Cao et al., 2010; Chrysikopoulos et al., 2010; Walsh et al., 2010) and predominantly affected by viral size rather than other factors (Dowd et al., 1998). For instance, coliphage MS2 (27–29 nm) are preferably to migrate (55%–79% of infiltration rate) than human adenoviruses (70–90 nm, 1%–31% of infiltration rate) (Wong et al., 2014; Kokkinos et al., 2015).

### 2.2. Morphological features

The morphological features of pathogens are another key influential factor on their migration (Ghanem et al., 2016). Bacterial morphological shapes include filamentous, helical, rod-shaped, ellipsoid and ovoid (Balkwill and Ghiorse, 1985). As rod-shaped colloids (width-length ratio = 1:6) are reported to have a slower migration rate than spherical ones (Ma et al., 2020), the elongated cells are more likely to adhere to particles than spherical ones (Feldner et al., 1983; Van Loosdrecht et al., 1987) and the length-width ratio is inversely proportional to bacterial migration rate (Salerno et al., 2006; Jiang and Bai, 2018). For instance, round bacterial cells have stronger migration ability comparing to elongated ones and >90% of the infiltrated bacteria from porous media have a width-length ratio >0.6 in a laboratory study (Weiss et al., 1995). Comparing to *Escherichia coli*, *Sphingopyxis alaskensis*, *Vibrio cholerae*, *Legionella pneumophila* and *Brevundimonas diminuta*, the slender spirillum-shaped *Hylemonella gracilis* showed a superior ability to pass through sterile membrane filters (Wang et al., 2008). Due to the increase of surface heterogeneity and hydrophobicity during the stretching process, the migration ability of rod-shaped bacteria is weaker than that of spherical ones (Seymour et al., 2013). Flagellum structure is reported to inhibit bacterial migration ability in porous media by promoting their motility and attachment efficiencies (Haznedaroglu et al., 2010).

Gram-negative bacteria with flagella are more difficult to migrate than non-movable ones (McClaine and Ford, 2002). In addition, flagella can promote the formation of biofilms and suppress bacterial migration (Abu-Lail and Camesano, 2003; Tong et al., 2010a; Du et al., 2020). Accordingly, in groundwater environment, common pathogenic bacteria with rod-shape and flagella-free are easy to migrate, e.g., *Streptococcus faecalis* and *Staphylococcus aureus*, whereas bacilli (*Escherichia coli*, *Pseudomonas aeruginosa*, *Mycobacterium tuberculosis*, *Mycobacterium leprosy*, *Bacillus cholerae*, *Mycobacterium typhoid*, etc.) are opposite (Nola et al., 2010).

For viruses, their main morphological shapes are categorized into spherical (e.g., phiX174, PM2, PRD1, phi6, MS2), filamentous (e.g., M13), polymorphic (e.g., MVL2) and tailed (e.g., T4, T7, AG3) (Dion et al., 2020). Similar to bacteria, spherical viruses are easier to migrate than polymorphic ones because of the directional effect that spherical viruses have a higher retention in porous media (Liu et al., 2010). Rod-like tailed phage reduces the migration by change the surface charge distribution during the transportation in porous media (Aronino et al., 2009). Spherical viruses are much easier to transport than non-spherical ones. For example, diffusivity of Q $\beta$  ( $16.6 \times 10^{-12} \text{ m}^2/\text{s}$ ) is higher than P22 ( $3.8 \times 10^{-12} \text{ m}^2/\text{s}$ ) (Baltus et al., 2017). Tailless viruses are reported to migrate more easily than tailed ones, and the migration abilities of viruses with different tail structure rank as the order of icosahedral Leviviridae [tailless,  $K_d = (2.6 \pm 1.2) \times 10^{-2}/\text{h}$ ] > Podoviridae [nonconstrictile short tail,  $K_d = (1.8 \pm 1.7) \times 10^{-2}/\text{h}$ ] > Siphoviridae [nonconstrictile long tail,  $K_d = (1.6 \pm 0.6) \times 10^{-2}/\text{h}$ ] (Ghanem et al., 2016). For instance, only 26% of tailed T4 viruses pass through the porous media comparing to tailless phage like phiX174 (48%) and MS2 (45%) (Aronino et al., 2009). In addition, viruses with spike proteins on surfaces, coronaviruses SARS-CoV-1 and SARS-CoV-2 as examples, have less mobility as they have higher surface roughness and can attach via receptor-specific interactions (Shen and Bradford, 2021; Shen et al., 2014). As common pathogenic viruses in groundwater (enterovirus, hepatitis virus, norovirus, rotavirus, and adenovirus) are spherical in shape, they are more likely to migrate in subsurface environment. Particularly for SARS-CoV-2 which are enveloped, spherical and tailless (Kumar et al., 2020), they are hypothetically to have a strong migration capability in porous media and pose significant risks.

### 2.3. Surface charge

Surface charge, mostly represented by zeta potential or isoelectric point, is also a key factor influencing pathogen migration by affecting the adsorption of pathogens in porous media (Pang et al., 2009). Normally, sand grains, clay minerals and bacterial pathogens like *Escherichia coli* O157:H7, *Yersinia enterocolitica* and *Enterococcus faecalis* are negatively charged in most pH range, and the electrostatic repulsion therefore inhibits pathogenic attachment and promotes their migration (Jacobs et al., 2007; Schinner et al., 2010). For instance, *Escherichia coli* with a lower surface charge (-44.7 mV, sticking efficiency = 0.02) are easier to migrate than those with higher surface charge (-22.1 mV, sticking efficiency = 0.4) (Lutterodt et al., 2009).

Surface charge density of pathogens can also influence their interactions with porous media and affect the migration behaviour (Baygents et al., 1998). A thinner double layer charge on the cell surface can increase isoelectric point and reduce the repulsive force between biocolloids and medium particles, benefiting pathogenic attachment on porous media and suppressing their migration (Zhang et al., 2018). For most RNA viruses, surface charge distribution is uneven (Jin and Flury, 2002). For example, caudophages have negatively charged head/tail tubes and positively charged filament (Penrod et al., 1996). Such uneven surface charge significantly affects the isoelectric point of tailed caudophages and

mitigates their transport in groundwater, further explaining their weaker migration ability than spherical viruses.

#### 2.4. Surface hydrophobicity

Pathogenic surface hydrophobicity is a crucial factor in pathogen migration process that hydrophobic pathogens normally have weaker migration abilities comparing with hydrophilic ones (Lutterodt et al., 2009; Liu et al., 2020). Short-term irreversible adsorption occurs between pathogen and medium surface through hydrophobic interaction (Gordesli and Abu-Lail, 2012; Zhao et al., 2014). Close to the medium surface, the hydrophobic interaction is detrimental to pathogen migration by eliminating the water between the interacting surfaces (Rodrigues et al., 2006; Zeraik and Nitschke, 2010), and thus pathogens with a weaker hydrophobicity migrate faster (Zhong et al., 2015). As most pathogenic outer membrane proteins are hydrophilic, pathogens with a higher content of outer membrane proteins have a stronger hydrophilicity and migration capability (Nikaido, 2003; Liu et al., 2020), whereas the hydrophobic functional groups on pathogenic surface could inhibit their migration (Alizadeh-Pasdar and Li-Chan, 2000). As the majority of subsurface media are hydrophilic (Cheng et al., 2012), we mainly consider their positive impacts on hydrophobic pathogen migration. For instance, with the decreasing aquifer saturation from 100% to 40%, the migration rates of hydrophobic pathogens decrease more significantly from 82.5% to 32.5% than hydrophilic pathogens (from 93% to 43.5%) (Gargiulo et al., 2008).

Generally, unenveloped and enveloped viruses are hydrophilic and hydrophobic (lipophilic), respectively (Feng et al., 2019). Owing to the dominance of both hydrophobic effects and electrostatic force amidst sorbent surface and capsid protein, enveloped viruses, such as mouse hepatitis virus and *Pseudomonas* phage  $\phi$  6, exhibit high adsorption potential on porous media with poorer migration ability (Ye et al., 2016). RNA-phages are more easily attached to media surface with their increasing hydrophobicity ( $GA\ 2300\ ng/cm^2 > Q\beta > MS2\ 2000\ ng/cm^2$ ) (Dika et al., 2013). Thus, Phage GA migrates slower in aquifer than phage MS2 (Armanious et al., 2016).

#### 2.5. Membrane structure and composition

Cell surface macromolecules including extracellular polymeric substances (EPS), outer membrane proteins and lipopolysaccharides affect the adsorption of pathogens and their migration behavior in porous media by altering cell surface charge and hydrophobicity (Tufenkji, 2007). Bacteria can be categorized as Gram-positive and Gram-negative ones according to the structure and composition of cell wall. Gram-positive bacteria have a large amount of peptidoglycan (about 40%–90% of dry cell weight) and some teichoic acids, exhibiting more negative charge and higher migration capability. In contrast, Gram-negative bacteria contain

11%–22% lipid, thin peptidoglycan layer (about 5% to 20% of the dry cell weight) and no teichoic acid on cell walls (Kang et al., 2006), commonly behaving relatively smaller migration ability. For example, *Staphylococcus aureus* showed a stronger migration ability (76.5% of retention) than *Escherichia coli* (99% of retention) (Weidhaas et al., 2014).

Bacterial EPS are reported to improve the adsorption efficiency (Liu et al., 2007; Tsuneda et al., 2003). The breakthrough concentration of *Escherichia coli* mutants with EPS production (55%) is less than the mutants not producing EPS (80%), indicating that EPS can promote microbial migration by inhibit the attachment (Tong et al., 2010b). In other studies, the outer membrane protein AG43 also promotes bacterial attachment and inhibits their migration by encouraging biofilm formation (Henderson, 1997; Klemm et al., 2003; Lutterodt et al., 2009). Lipopolysaccharides on cell surface can decelerate bacterial migration by promoting their adsorption on media (Abu-Lail and Camesano, 2003; Walker et al., 2004).

Different from pathogenic bacteria, viruses have unique structure and are categorized into enveloped, unenveloped (Blanco et al., 2019) and vesicle ones (Zhang et al., 2021d). Previous studies mainly focus on the migration of non-enveloped viruses. Enveloped viruses have similar lipid bilayer membrane structure like bacteria, except for viral glycoproteins which are more sensitive to environmental changes (Wolfe et al., 2017). As viral envelope is hydrophobic and more prone to adsorb on organic particles or solid surfaces, their migration rates are relatively low (Gundy et al., 2009; Paul et al., 2021).

#### 2.6. Evaluation of pathogen migration capacity by physiological properties

Pathogen size, morphological features, surface charge, surface hydrophobicity and membrane structure/composition are the four key influential variables affecting pathogen migration abilities in subsurface environment. Nevertheless, most previous studies only focus on limited factors or pathogens, and there is lack of assessment algorithm to evaluate the migration ability of a variety of pathogens in subsurface environment (Dika et al., 2015; Blanco et al., 2019; Paul et al., 2021). From the effects of physiological properties on pathogen migration given in the literature, pathogen size is the most important variable, followed by surface charge represented by isoelectric point, surface hydrophobicity and membrane structure/composition represented by contact angle, and morphological features of width-length ratio and flagella (Haznedaroglu et al., 2010; Ghanem et al., 2016; Pang et al., 2021). Based on the importance of each factor reported by previous literatures, we propose an assessment algorithm considering the five indices to quantitatively evaluate the migration ability of pathogens, including pathogen size (40%), isoelectric point (20%), contact angle (20%), width-length ratio (15%) and flagella (5%) (Table 1). It is worth noting that the importance of these factors

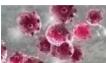
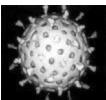
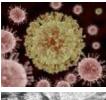
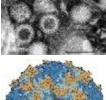
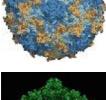
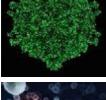
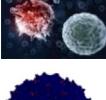
**Table 1**  
Criteria for estimating the migration ability of pathogens.

Weighted term	Parameter	Weight	Classification and scores				
Pathogen size	Size ( $\mu\text{m}$ )	40%	<b>&lt;0.06</b>	<b>0.06–0.30</b>	<b>0.30–1.00</b>	<b>1.00–3.00</b>	<b>&gt;3.00</b>
			1.0	0.8	0.6	0.4	0.2
Surface charge	Isoelectric point	20%	<b>&lt;4.0</b>	<b>4.0–5.0</b>	<b>5.0–6.0</b>	<b>&gt;6.0</b>	–
			1.0	0.7	0.4	0.1	–
Surface hydrophobicity	Contact angle	20%	<b>&lt;25°</b>	<b>25°–35°</b>	<b>35°–45°</b>	<b>45°–65°</b>	<b>&gt;65°</b>
			1.0	0.9	0.7	0.5	0.1
Morphological features	Width-length ratio	15%	<b>&gt;0.8</b>	<b>0.8–0.6</b>	<b>0.6–0.4</b>	<b>0.4–0.2</b>	<b>&lt;0.2</b>
			1.0	0.8	0.6	0.4	0.2
Membrane structure	Flagella	5%	<b>No</b>	<b>Yes</b>			
			1.0	0.5			

The bold fonts indicate criteria for classification, and normal fonts indicate scores.

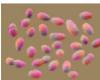
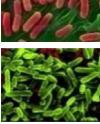
**Table 2**

Estimated migration abilities of different pathogens based on their size, morphological features, surface charge, surface hydrophobicity and membrane structure<sup>1</sup>.

Pathogens	Morphology <sup>2</sup>	Flagella	Size (nm <sup>2</sup> )	pH <sub>PZC</sub>	Contact angle <sup>3</sup>	Reference	Estimated migration ability	
Non-enveloped viruses	Norovirus		Spherical (1)	No	38	5.0	42°	(Redman et al., 1997; Collins et al., 2006) 0.88
	Adenovirus		Spherical (1)	No	70–90	7.2	68°	(Flint, 2004; Baker et al., 2021; Dang and Tarabara, 2021) 0.56
	Rotavirus		Spherical (1)	No	80	3.9	-	(Collins et al., 2006; Jimenez-Zaragoza et al., 2018) 0.86
	Rhinovirus		Spherical (1)	No	28–30	6.4	-	(Flint et al., 2004) (Flint et al., 2004) 0.72
	Reovirus		Spherical (1)	No	70–80	3.9	-	0.82
	Poliovirus		Spherical (1)	No	28–30	3.8–8.2 (6.0)	PRD1 > MS2(33°)≈ Echovirus > Coxsackievirus > Poliovirus > T2 ≈ T4 > ΦX174(26°)	(Schijven and Hassanizadeh, 2000; Flint et al., 2004; Collins et al., 2006) 0.86
	Echovirus		Spherical (1)	No	28–30	5.1–6.4 (5.8)	(Flint et al., 2004)	0.86
	Coxsackievirus		Spherical (1)	No	28–30	4.8–6.1 (5.4)	(Flint et al., 2004)	0.86
Phage viruses	MS2		Spherical (1)	No	26–27	3.9	(Shields and Farrah, 1983; Penrod et al., 1996; Schijven and Hassanizadeh, 2000; Gitis et al., 2002; Pang et al., 2009; Chrysikopoulos and Syngouna, 2012a)	0.98
	PRD-1		Spherical (1)	No	62	3–4.2	(Schijven and Hassanizadeh, 2000; Abudalo et al., 2005; Collins et al., 2006;)	0.86
	Φx174		Spherical (1)	No	25–27	6.0	(Schijven and Hassanizadeh, 2000; Collins et al., 2006; Zhang et al., 2010; Chrysikopoulos and Syngouna, 2012b)	0.86

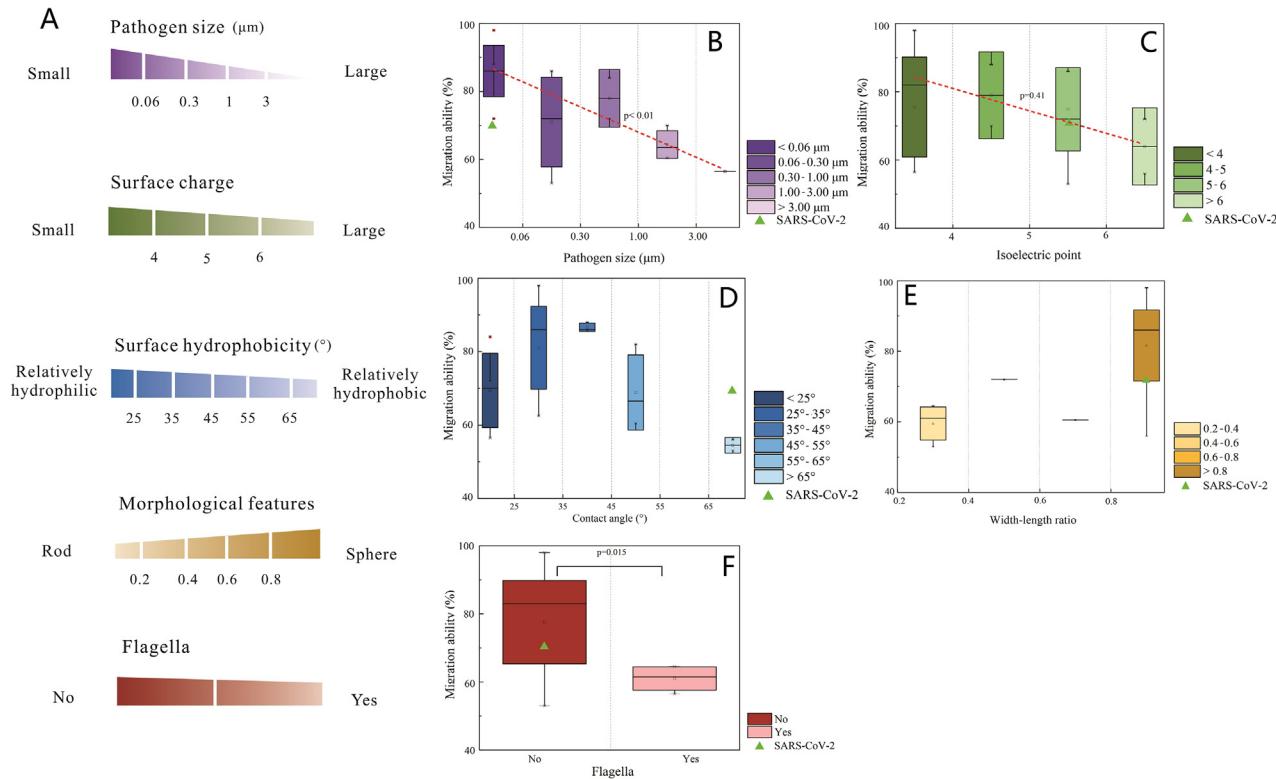
(continued on next page)

**Table 2 (continued)**

Pathogens	Morphology <sup>2</sup>	Flagella	Size (nm <sup>2</sup> )	pH <sub>PZC</sub>	Contact angle <sup>3</sup>	Reference	Estimated migration ability		
T2 phage		Tailed (0.524)	No	Head 109, tail 99	4.2	(Deblois and Wesley, 1977)	0.72		
T4 phage		Tailed (0.479)	No	Head 84–100, tail 100	4–5	(Carrascosa and Kellenberger, 1978)	0.72		
Phage λ		Tailed (0.265)	No	Head 54, tail 150	3.9–5.0	–	(Penrod et al., 1996)	0.61	
<i>E. coli</i> phage		Tailed (0.331)	No	Head 53, tail 107	4.37	66.8°	(Yu, 2016; Qin, 2020)	0.53	
Pathogenic viruses (Enveloped virus)	SARS-CoV-2		Spherical (1)	No	60–140	5.9	>60°	(Pandey, 2020; Zhu et al., 2020)	0.70
Bacteria	<i>Enterococcus faecalis</i>		Spherical or chain (1)	No	1290–1380	4.6	19°	(Fernando et al., 1991; Millsap et al., 1994; Signoretto et al., 2000; Lu Ting, 2012)	0.70
	<i>Streptococcus</i>		Spherical or chain (1)	No	600–1000	5.3	21°	(Millsap et al., 1994; Kamezawa et al., 1997)	0.72
	<i>Bacillus subtilis</i>		Oval to cylindrical (0.3)	Yes	(700–800) × (2000–3000)	3.5	32°	(Ahimou et al., 2001; Ren et al., 2018; Wu, 2021; Rong, 2008)	0.63
	<i>Clostridium perfringens</i>		Fusiform (0.657)	Yes	(300–2000) × (1500–2000)	4.0	–	(Ando and Tsuzuki, 1984)	0.61
	<i>Staphylococcus aureus</i>		Spherical (1)	No	800	4.0	19.5°	(Rawlinson et al., 2011; Zmantar et al., 2011)	0.84
	<i>Escherichia coli</i>		Rod (0.35)	Yes	2000 × 700	2.0–4.3	22°	(Gilbert et al., 1991; Lytle et al., 1999; Oh et al., 2007; Pang et al., 2009; Wu, 2021)	0.65
	<i>Pseudomonas putida</i>		Rod (0.231)	Yes	(500–1000) × (1500–5000)	3.5	24°	(Abu-Lail and Camesano, 2003; Rong, 2008)	0.57

Notes:

<sup>1</sup>Detailed calculation and estimated data for each physiological character are listed in Supplementary data Table S1.<sup>2</sup>The number in brackets is the width-length ratio of pathogenic in the column of morphology.<sup>3</sup>“–” indicates not retrieved from the literature.



**Fig. 1.** (A) Impacts of pathogenic physiological characteristics on the estimated migration abilities of pathogens. (B) The estimated migration ability vs pathogen size. (C) The estimated migration ability vs surface charge (isoelectric point). (D) The estimated migration ability vs surface hydrophobicity (contact angle). (E) The estimated migration ability vs morphological features (width-length ratio). (F) The estimated migration ability vs membrane structure (flagella).

depends on the geochemical and hydrological conditions in subsurface system, and this assessment algorithm is only suitable for most scenarios.

For the classification of pathogen size, viruses ( $0.023\text{--}0.08\ \mu\text{m}$ ) generally have stronger migration ability than bacteria ( $0.5\text{--}3\ \mu\text{m}$ ) (Walshe et al., 2010), and a significant stronger migration ability is observed for viruses with size  $<0.06\ \mu\text{m}$  (Cao et al., 2010; Chrysikopoulos et al., 2010; Walshe et al., 2010). Accordingly, the size classification includes  $<0.06$ ,  $0.06\text{--}0.30$ ,  $0.30\text{--}1.00$ ,  $1.00\text{--}3.00$  and  $>3.00\ \mu\text{m}$ . Isoelectric point has been widely studied in migration of many pathogens, and the four classification groups are set as  $<4.0$ ,  $4.0\text{--}5.0$ ,  $5.0\text{--}6.0$  and  $>6.0$  based on the normal isoelectric point range of pathogens. The contact angle is directly proportional to the hydrophobicity and inversely proportional to migration ability. As the hydrophobic interaction between surfaces is effective at contact angle  $>65^{\circ}$  and some published contact angles of pathogens are as follows: PRD1 > MS2 ( $33^{\circ} \approx$  Echovirus > Coxsackie virus > Poliovirus > T2  $\approx$  T4 >  $\Phi$ X174  $\approx$   $26^{\circ}$ ) (Dika et al., 2015; Shields, 1986), the classification of contact angle is  $<25^{\circ}$ ,  $25^{\circ}\text{--}35^{\circ}$ ,  $35^{\circ}\text{--}45^{\circ}$ ,  $45^{\circ}\text{--}65^{\circ}$  and  $>65^{\circ}$ . The width-length ratio representing pathogenic morphological characteristics increases from 0.0 (rod-shaped) to 1.0 (spherical-shaped), calculated as the ratio of head width to total length of tail phage and classified into five categories of  $>0.8$ ,  $0.8\text{--}0.6$ ,  $0.6\text{--}0.4$ ,  $0.4\text{--}0.2$  and  $<0.2$ .

Based on the assessment algorithm and classification, the migration abilities of many well-known pathogens are calculated and listed in Table 2. For instance, *Escherichia coli* has a low score of 0.65, indicating its relatively poor migration ability, owing to the large size (about  $2000\ \text{nm} \times 700\ \text{nm}$ ) and non-spherical shape (rod-shaped) (Pang et al., 2009). In contrast, *Staphylococcus aureus* has a higher score of 0.84, indicating a stronger migration ability, owing to the smaller size (about  $800\ \text{nm}$ ) and spherical shape

(Rawlinson et al., 2011). The estimated average migration abilities of Gram-positive bacteria (0.7, including *Enterococcus faecalis*, *Streptococcus*, *Bacillus subtilis*, *Clostridium perfringens* and *Staphylococcus aureus*) are stronger than those of Gram-negative bacteria (0.61, including *Escherichia coli*, *Pseudomonas putida*). For viruses, *Norovirus* has a small size (about  $0.035\ \mu\text{m}$ ), non-membrane and spherical shape (Collins et al., 2006), resulting in a relatively higher score of 0.88. Generally, most viruses except for some tailed phages have a higher score (0.86–0.98) than bacterial pathogens, suggesting higher migration abilities of viruses in subsurface environment, consistent with some previous laboratory studies (Bradford et al., 2006a; da Silva et al., 2011).

From the calculated migration abilities in Table 2, the correlations between pathogen migration and individual physiological characteristics are illustrated in Fig. 1. From the general trend in Fig. 1A, pathogenic migration abilities dramatically decrease with pathogen size, but only slightly decrease with surface charge, surface hydrophobicity and morphological structure. Regression curve in Fig. 1B also evidences that pathogen size is inversely proportional to migration ability. As for surface charge (Fig. 1C), surface hydrophobicity (Fig. 1D) and morphological structure (Fig. 1E), the correlations are not significant and suggest that pathogen size is the most key factor influencing pathogen migration in subsurface environment. It might be explained by limited experimental data till now on these physiological parameters, and only pathogen size is investigated in most previous studies. It is worth highlighting that, although the weight of flagella is only 5% in the assessment algorithm, pathogens with and without flagella exhibit significant difference (Fig. 1F), hinting that flagellum is also a key factor affecting pathogen behavior in aquifers.

As an enveloped (Ke et al., 2020), spherical and pleomorphic virus (Zhu et al., 2020), SARS-CoV-2 virus has a size of

60–140 nm. SARS-CoV-2 consists of four major structural proteins: spike protein (S), membrane protein (M), nucleocapsid protein (N) and envelope protein (E) (Sarkar and Saha, 2020). Among them, spike protein promotes host attachment and viral cell membrane fusion for infection (Wu et al., 2020), and adheres to other materials (Pandey, 2020). Although the interfacial dynamics of SARS-CoV-2 pseudoviruses were investigated (Liu et al., 2021b), its migration ability is still not clear. Here, we speculate that the adhesion between SARS-CoV-2 and porous media is mainly attributing to spike protein. Although the isoelectric point and hydrophobicity of SARS-CoV-2 remain unclear (Scheller et al., 2020), it can be estimated by the ProtParam tool of ExPASy (Gasteiger et al., 2003) or of the same value of 5.90 as spike protein (Pandey, 2020). In addition, as both spike protein and capsule lipid bilayer are hydrophobic (Pandey, 2020), SARS-CoV-2 is highly possible hydrophobic. Based on this information, the migration ability of SARS-CoV-2 is scored as 0.70 in the proposed assessment algorithm, indicating that SARS-CoV-2 has a relatively strong migration ability in subsurface environment.

### 3. Environmental geochemical and hydrological variables

Although pathogen migration behavior is intrinsically determined by their physiological characteristics, their migration process in subsurface environment is also influenced by geochemical variables of porous media and hydrological variables in groundwater. Considering the key variables that associate with pathogenic physiological characteristics, those environmental geochemical and hydrological variables include soil structure, groundwater chemical composition, pH, ionic strength, temperature, moisture, and hydrodynamic condition.

#### 3.1. Soil texture

Soil texture determines soil pore channels which directly affects pathogen migration abilities in subsurface environment, and the key factors include grain size, surface roughness and heterogeneity (Sasidharan et al., 2017a). Pathogen migration in porous media follows filtration mechanisms, e.g., strain and deposition. Bigger grain size increases pathogen migration (Bai et al., 2016; Mohanty and Boehm, 2014), evidenced by their faster migration in quartz sands with a larger size (0.71 mm) than smaller one (0.15 mm) (Bradford et al., 2006a). For grain size <50 μm, pathogen migration is suppressed by straining (Torkzaban et al., 2015).

Surface roughness is proportional to the specific surface area of porous media. Higher surface roughness and specific surface area increases the active attachment sites to capture pathogens and inhibits their migration (Morales et al., 2009; Krishna Darbha et al., 2012; Torkzaban and Bradford, 2016). For instance, crude biochar has rougher surface with the specific surface area 5 orders of magnitude larger than sand, resulting in a significantly stronger adsorption of *Escherichia coli* through biochar (log removal of 2.32) than sand (log removal of 0.29) (Mohanty et al., 2014). In addition, the adsorption of *Escherichia coli* are equal to or higher on rough beads than smooth beads in a simulated groundwater migration system (Shellenberger and Logan, 2002). Moreover, nanoscale surface roughness can alter the interaction energies by decreasing the primary minimum depth and decrease pathogen attachment on medium surface (Shen et al., 2018; Rasmussen et al., 2019), although secondary energy minimum and wells might be increased under some unfavorable chemical conditions, e.g., low ionic strength (Shen et al., 2012). Taking these two mechanisms together, surface roughness can promote pathogen deposition and inhibit their migration.

The heterogeneity of porous media can affect pathogen migration by altering the collision efficiencies and deposition capacities,

because of the higher attachment at secondary minima at concave locations or detachment from primary minima at nanoscale convex asperities (Shen et al., 2020). Fine-textured and poor-structured soils intercept bacteria through mechanical filtration for their small pore sizes and low hydraulic conductivity (Morales et al., 2015). Accordingly, an increasing virus migration is observed with the standard deviations of hydraulic conductivity in heterogeneous aquifers in a migration model (Torkzaban et al., 2019).

#### 3.2. Groundwater chemical composition

Groundwater chemical composition determines pathogen migration through hydrophobic and hydrophilic interactions in groundwater, thus linked to pathogen surface hydrophobicity. A wide variety of inorganic and organic materials exist as colloids in groundwater, including mineral precipitates like iron and manganese oxides, rock and mineral fragments and macromolecular components of natural organic matters (McCarthy and McKay, 2004; Foppen et al., 2008; Walshe et al., 2010), exhibiting strong interactions and changing pathogenic migration behavior in groundwater (Foppen et al., 2008; Yang et al., 2012b; Cai et al., 2013). The presence of phosphate can alter bacterial EPS and increase the repulsion between pathogens and media. For instance, phosphate is reported to promote the migration of *Escherichia coli* O157:H7 in quartz sand (Wang et al., 2011). Another study finds that silicate could increase bacterial surface charge and weaken the electrostatic repulsion between bacteria and media, thereby reducing the migration ability of *Escherichia coli* (Dong et al., 2014). Colloid-facilitated transport is an important mechanism of pathogen migration, and it is believed that colloids could increase the migration velocity and distances of pathogens (Babich and Stotzky, 1980; Jin et al., 2000). Nevertheless, organic and inorganic colloids exhibit entirely different transport mechanisms (Katzourakis and Chrysikopoulos, 2014; Bellou et al., 2015; Syngouna and Chrysikopoulos, 2016). Organic colloids would form stable aggregates with pathogens via interparticle interaction and expedite their migration owing to the rough surface with functional groups and long-chain macromolecular structure (Akbour et al., 2002). In contrast, the inorganic colloids are prone to agglomerate causing blockage and inhibiting pathogen migration (Qin et al., 2020). In porous media, the migration of *Escherichia coli* phage vB\_EcoM-ep3 with silica colloids (deposition rate of  $7.23 \times 10^{-2} \text{ s}^{-1}$ ) is faster than with fusiform humic acid colloids (deposition rate of  $5.14 \times 10^{-3} \text{ s}^{-1}$ ) (Qin, 2020). Another study reports 27.73% increase of viral migration rate in the presence of humic acid colloids (Yu, 2016).

Natural dissolved organic matters also influence the deposition and migration behavior of pathogens (Foppen et al., 2008; Yang et al., 2012a; Zhao et al., 2014). As they can be adsorbed on hydrophobic organic compounds and increase solubility (Foppen et al., 2008), natural dissolved organic matters might also interact with pathogens with high hydrophobicity and accelerate their migration. For instance, the penetration rate of *E. coli* increases about 20% in the presence dissolved humic acids (Foppen et al., 2008). In addition, dissolved organic matters can also inhibit microbial attachment on porous media via competitive adsorption by occupying the active sites on medium and increasing the steric hindrance of deposition, encouraging pathogen transportation (Foppen et al., 2006). Some evidence can be found from previous studies that bacterial migration is accelerated with the declined adsorption of hematite onto cell surfaces in the presence of humic acids (Foppen et al., 2008; Yang et al., 2016). Mineral-associated organic matters are reported to significantly promote the migration of phage MS-2 in phosphate-buffered saline (Zhuang and Jin, 2003a). Generally, pathogens migrate faster in organic-rich aquifers which require extensive attentions.

### 3.3. pH

pH together with pathogenic isoelectronic point determines pathogen migration in subsurface environment by changing the adsorption–desorption and diffusion process. Pathogen adsorption on porous media is enhanced at lower pH, which decreases their migration rates (Zhang et al., 2018). At higher pH value, pathogenic surface is more negatively charged and is harder to be adsorbed by aquifer media. Some experimental evidence shows that MS2 bacteriophage has a higher migration rate in slightly alkaline groundwater (1.13 m/s, pH = 8.1) than acidic groundwater (0.90 m/s, pH = 6.1) (Schulze-Makuch et al., 2003), and the diffusion coefficients of viruses increase with pH value (Langlet et al., 2008). Elevated solution pH is reported to increase bacterial migration (Kim et al., 2009), whereas the declined pH enhances virus attachment to aquifer media and colloids, resulting in more retarded virus migration with lower peak-concentration, longer peak-concentration time and higher apparent collision efficiency (Walshe et al., 2010; Zhang et al., 2018). More bacteriophages MS2 and  $\Phi$ X174 are attached on the solid-water interface under low pH condition and their migration is slowed (Torkzaban et al., 2006). In addition, higher pH can also enhance pathogen migration by encouraging the deprotonation of functional groups on cell surface and increasing the electrostatic repulsion (Schinner et al., 2010; Zhang et al., 2018). It is also worth noting that, when pH is close to the isoelectric point of pathogens, viruses might aggregate due to the weak electrostatic repulsion interactions, possibly causing blockage and immobile (Schinner et al., 2010; He et al., 2014). Generally, pathogen migration in subsurface environment is faster under neutral or weak alkali conditions, deserving more attentions.

### 3.4. Ionic strength

Ionic strength is reported to affect pathogenic attachment efficiency and then their migration ability (Zhuang and Jin, 2003b), explained as the change of pathogenic surface charge and membrane composition by ionic strength (Degrève et al., 2012). Ionic strength has a remarkable influence on the electrokinetic properties of either pathogens or grains (Kim and Walker, 2009) by decreasing the repulsive energy barrier and increasing secondary minima (Li et al., 2020). As viruses have lower attachment efficiencies on the negatively charged quartz sand at lower ionic strength (Gomez-Suarez et al., 2001), higher ionic strength increases pathogenic attachment efficiency (Han et al., 2006; Alexis et al., 2007; Rong et al., 2007; Kim and Walker, 2009). The electrostatic repulsion is reported to reduce as ionic strength increases, promoting the adhesion of bacterial proteins (Cui et al., 2015; Miao et al., 2015), and the migration rate of human adenovirus decreases with the increasing ionic strength from 0.001 mol/L to 0.1 mol/L (Wong et al., 2014). In addition, ionic valence is also a key factor affecting pathogen attachment and polyvalent cations are more likely to slow pathogen migration (Kim et al., 2009; Schinner et al., 2010). For instance, bacteria have higher adhesion efficiencies in  $\text{CaCl}_2$  solution than KCl solution (Chen and Walker, 2007). Generally, pathogens have relatively high migration rates at the ionic strength of groundwater (about 0.02–0.04 mol/L).

Ionic strength can also change pathogen migration by altering the molecular structure of membrane proteins. By studying two noroviruses (GI.1 and GII.4) with the same isoelectric point but different membrane structure and composition, ionic strength shows distinct effects on their attachment on media (positive correlation with GII.4 but negative correlation with GI.1) (da Silva et al., 2011), explained by the significant shifts from  $\beta$ -strand to  $\alpha$ -helix and slight losses of ordered secondary structure in GI.1 and barely change in GII.4 (Samandoulgou et al., 2015).

### 3.5. Temperature

A rise in temperature can decrease the energy barrier and water viscosity, thus increasing pathogen attachment and inhibiting their migration (Gharabaghi et al., 2015; Sasidharan et al., 2017b). Many studies have reported that temperature influences viral attachment to solid surfaces and affects their transport in subsurface environment (Gallardo-Moreno et al., 2003; Bradford et al., 2006b; Castro and Tufenkji, 2007; Kim and Walker, 2009; Chrysikopoulos and Aravantinou, 2014). The attachment rate coefficient of viruses PRD1 and  $\Phi$ X174 increases up to 109% when temperature increases from 4 °C to 20 °C (Sasidharan et al., 2017b). Additionally, pathogenic physiological properties also change with temperature, e.g., EPS formation, protein folding and hydrophobicity, resulting in an increasing attachment of pathogens on some adsorbents with temperature (Bales et al., 1991; Bellamy et al., 1985). As groundwater temperature normally ranges of 10–22 °C in the thermostatic zone and deep phreatic water rises 1 °C for every 33 m increase of depth (Yates et al., 1985; Gunawardhana et al., 2009; Kar et al., 2010; Vanderzalm et al., 2010), pathogen migration in deep groundwater is relatively slow, but it is non-neglectable fast in shallow groundwater and poses significant risks for drinking water safety.

### 3.6. Moisture

Pathogen transport in subsurface environment is governed by advection, dispersion, and inactivation across multiple interfaces. As pathogens mainly enter the aquifer through the vadose zone, their vertical migration in the vadose zone is critical and moisture is a key factor (Flury and Aramrak, 2017; Kamrani et al., 2018; Sirivithayapakorn and Keller, 2003). Pathogen migration in soils of the vadose zone is inhibited due to the increasing of attachment as the moisture decreases (Gargiulo et al., 2008; Kim et al., 2008). As pathogens and the air–water interface are both negatively charged (Graciaa et al., 1995; Marinova et al., 1996), pathogens are usually electrostatically repelled from the air–water interface and attach to surfaces by thin water films and capillary menisci at the air–water–solid interface. Accordingly, viral attachment on the solid–water interface is more dominant than air–water interface, and it is controlled by solution chemistry, particle surface charge, and hydrophobicity (Wan and Tokunaga, 2002). In the vadose zone with lower saturation conditions, pathogens can be retained significantly by their attachment to the solid–water interface due to electrostatic interactions. As moisture decreases, viral attachment to solid–water and air–water–solid interface are enhanced (Flury and Aramrak, 2017), thus slowing pathogen vertical migration. For instance, the retention of viruses MS2 and  $\Phi$ 174 increases as moisture decreases in laboratory test (Torkzaban et al., 2006). However, the presence of colloids in the vadose zone can enhance pathogen irreversible attachment to solid–water interface and adsorption to air–water–solid interface, mitigating the effects of moisture (Chu et al., 2001). For example, silica colloids are reported to inhibit the migration of *Escherichia coli* phage in saturated zone but oppositely promote their migration in the vadose zone when moisture decreases (Qin et al., 2020). Therefore, the effects of moisture on migration or remigration ability of pathogens during the intermittent water infiltration deserve attentions.

### 3.7. Hydrodynamic condition

Besides geochemical variables, pathogen migration in subsurface environment is also related to the hydrogeological conditions, such as flow rate (Yan et al., 2020). Lower flow rates and interruptions with higher hydraulic residence time can retard *Escherichia coli* and slow their migration (Sasidharan et al., 2017a). Conversely, a higher flow rate is reported to promote pathogen migration

(Predelus et al., 2017). The positive correlation between pathogen migration ability and flow rate can be firstly explained by the increasing higher hydraulic residence time of pathogens in porous media (Meinders et al., 1994; Xu et al., 2005). Secondly, the deposition kinetics are determined by the adhesive strength, which is higher and benefits pathogen deposition on solid surface at higher flow rates (Ko and Elimelech, 2000; Xu and Logan, 2006; Sasidharan et al., 2017a). Last but not least, lower flow rate means weaker hydrodynamic forces acting on pathogens which satisfy the torque balance on the surface of aquifer media (Bradford et al., 2011; Sasidharan et al., 2017a).

### 3.8. Subsurface environment with high risks of pathogen migration

Considering these key environmental geochemical and hydrological variables, the risks of pathogen migration in subsurface can also be evaluated by our proposed assessment algorithm but with different indices, following a similar rating and weighting

scheme of DRASTIC suggested by the United States Environmental Protection Agency (USEPA) (Aller et al., 1987; Barbulescu, 2020). Based on the results of previous studies (Kortelainen and Karhu, 2009; Zhang et al., 2018, 2021e; Arcega-Cabrera et al., 2021; White et al., 2021), the five indices include moisture (25%), flow rate (25%), surface roughness of medium (25%), ionic strength (15%) and pH (10%) (Table 3). Again, this assessment algorithm is only suitable for most subsurface environment, and the importance of these geochemical and hydrological factors depends on pathogenic physiological features and vary significantly across species.

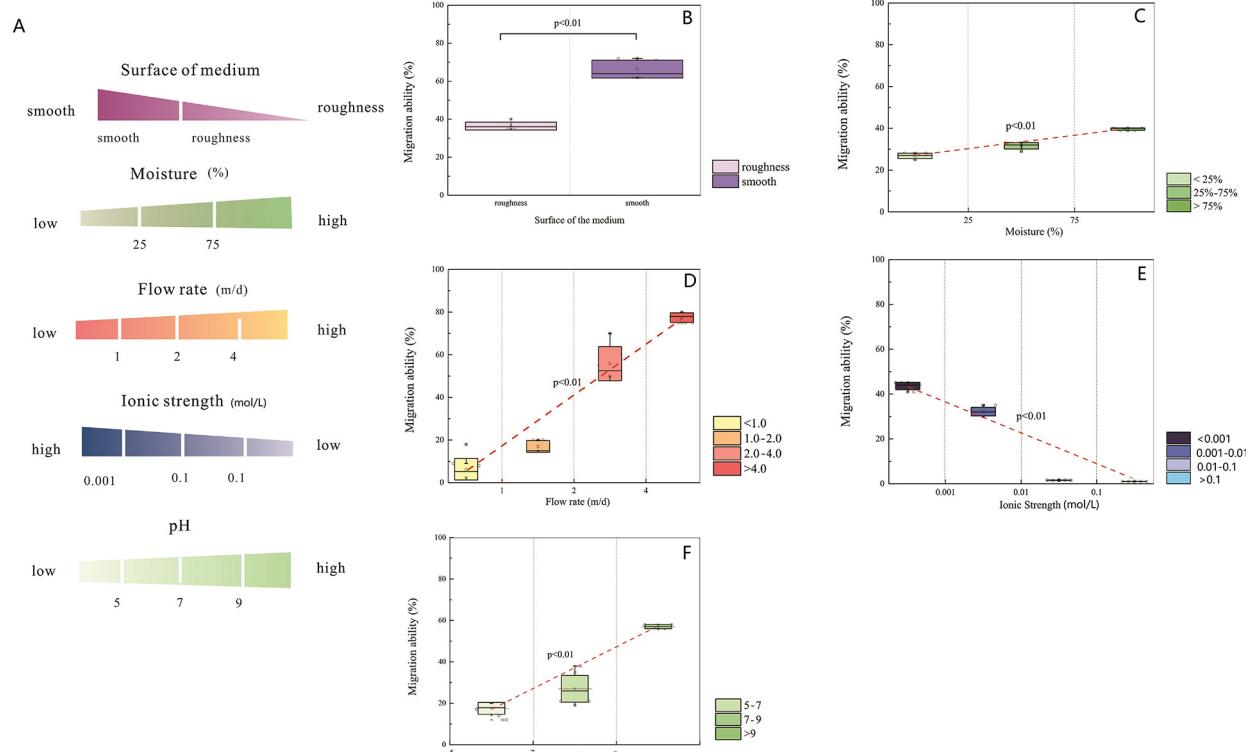
For the classification of moisture, it ranges from 25% in the vadose zones to 82% of saturated water holding capacity in the aquifers (Gargiulo et al., 2008), and thus divided into three categories of <25%, 25%–75% and >75%. Considering the normal groundwater velocity, the four classification groups of flow rate are set as <1, 1–2, 2–4 and >4 m/d. As the ionic strength of groundwater is normally <0.1 mol/L (Wong et al., 2014), it is categorized

**Table 3**

Scoring criteria of migration ability of pathogen with different environmental conditions.

Scoring item	Weight	Classification and scores		
Surface roughness	25%	<b>Roughness</b>	<b>Smooth</b>	
		1.0	5.0	
Moisture	25%	<b>&gt;75%</b>	<b>25%–75%</b>	<b>&lt;25%</b>
		1.0	0.7	0.1
Flow rate (m/d)	25%	<b>&gt;4</b>	<b>2–4</b>	<b>1–2</b>
		1.0	0.7	0.4
Ionic strength (mol/L)	15%	<b>&gt;0.1</b>	<b>0.01–0.1</b>	<b>0.001–0.01</b>
		1.0	0.8	0.5
pH	10%	<b>&gt;9.0</b>	<b>7.0–9.0</b>	<b>&lt;7.0</b>
		1.0	0.7	0.3

The bold fonts indicate criteria for classification, and normal fonts indicate scores.



**Fig. 2.** (A) Impacts of geochemical and hydrological variables in subsurface environment on the estimated migration abilities of pathogens. (B) The estimated migration ability vs medium surface. (C) The estimated migration ability vs moisture. (D) The estimated migration ability vs flow rate. (E) The estimated migration ability vs ionic strength. (F) The estimated migration ability vs pH. Data are from (Walsh et al., 2010; Wong et al., 2014; Lv et al., 2016; Sasidharan et al., 2016; Yu, 2016; Li, 2018; Zhang et al., 2018).

into four groups of <0.001, 0.001–0.01, 0.01–0.1 and >0.1 mol/L. Generally, the pH value in groundwater ranges from 5.0 to 9.0 (Schulze-Makuch et al., 2003), which are then evenly divided into three categories of <7.0, 7.0–9.0 and >9.0.

Taking bacteriophage MS2 as a typical and representative pathogen, we evaluate its migration abilities under different environmental geochemical and hydrological conditions (Fig. 2). All variables show significant influence on the migration of bacteriophage MS2 (Fig. 2A). The migration ability is significantly higher in media with smooth surfaces than rough ones ( $p < 0.01$ . Fig. 2B). For other variables, remarkable positive correlations are found between MS2 migration ability and moisture (Fig. 2C), flow rate (Fig. 2D) and pH (Fig. 2F), whereas ionic strength exhibits a negative relationship with migration ability (Fig. 2E). Accordingly, the porous media in subsurface environment with smooth surface, high moisture, high flow rate, high pH and low ionic strength are of significantly high biosafety risks regarding pathogen migration, e.g., coarse sandy, karst and fissure aquifers. As SARS-CoV-2 has a considerable migration ability, their spreading potential should deserve careful attentions in these areas with biosafety fragility.

#### 4. Conclusion and prospective

The outbreak of COVID-19 and other emerging diseases has aroused world-wide concerns about pathogen migration and spread in natural environment. As the main source of drinking water, groundwater is threatened by pathogen vertical migration from medical waste disposal, agricultural activities, as well as landfills, and might behave as a secondary source for the concealed pathogen spread, urgently requiring explicit studies on the influential factors on pathogen migration in subsurface system. In this work, we comprehensively review physiological characteristics, geochemical properties and hydrological variables influencing pathogen migration in subsurface environment. Five pathogenic physiological characteristics including pathogen size, isoelectric point, contact angle, width-length ratio and flagella intrinsically determine pathogen migration ability; the deterministic geochemical properties and hydrological variables include moisture, flow rate, surface roughness, ionic strength and pH. It is obvious that pathogen migration is complicated and affected by many intrinsic and extrinsic factors in subsurface environment, challenging appropriate evaluation of their migration abilities, particularly for those pathogens without sufficient information. Here, we propose an assessment algorithm and two rating/weighting schemes for pathogenic physiological characteristics and geochemical/hydrological variables to evaluate the migration abilities and risks of pathogens in subsurface system. In addition, we use this method to predict the migration abilities of SARS-CoV-2 viruses and subsurface systems with biosafety fragility. Although these algorithms and schemes are still an exploratory attempt owing to the limited works studying pathogen migration currently and different investigated parameters across literatures, our work highlights and points out the prior variables to explore in future experimental and field studies. The non-neglectable biosafety risks in subsurface system call for our attentions on both receiving pathogens by above-ground natural human disturbance (rainfall infiltration, leaching, artificial groundwater recharge, etc.) via vertical migration as a sink and spreading pathogens to nearby groundwater resources via horizontal migration as a source.

#### Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

#### Acknowledgement

This work is financially supported by the National Science Foundation of China (Grant No. 52091543). DZ also acknowledges the support of Chinese Government's Thousand Talents Plan for Young Professionals.

#### Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.gsf.2021.101346>.

#### References

- Abu-Lail, N.I., Camesano, T.A., 2003. Role of lipopolysaccharides in the adhesion, retention, and transport of *Escherichia coli* JM109. *Environ. Sci. Technol.* 37 (10), 2173–2183.
- Abudalo, R.A., Bogatsu, Y.G., Ryan, J.N., Harvey, R.W., Metge, D.W., Elimelech, M., 2005. Effect of ferric oxyhydroxide grain coatings on the transport of bacteriophage PRD1 and *Cryptosporidium parvum* oocysts in saturated porous media. *Environ. Sci. Technol.* 39 (17), 6412–6419.
- Ahimou, F., Paquot, M., Jacques, P., Thonart, P., Rouxhet, P.G., 2001. Influence of electrical properties on the evaluation of the surface hydrophobicity of *Bacillus subtilis*. *J. Microbiol. Methods* 45 (2), 119–126.
- Ahmed, W., Angel, N., Edson, J., Bibby, K., Mueller, J.F., 2020. First confirmed detection of SARS-CoV-2 in untreated wastewater in Australia: A proof of concept for the wastewater surveillance of COVID-19 in the community. *Sci. Total Environ.* 728, 138764.
- Akbour, R.A., Douch, J., Hamdani, M., Schmitz, P., 2002. Transport of kaolinite colloids through quartz sand: Influence of humic acid,  $\text{Ca}^{2+}$ , and trace metals. *J. Colloid Interface Sci.* 253 (1), 1–8.
- Alegbeleye, O.O., Sant'Ana, A.S., 2020. Manure-borne pathogens as an important source of water contamination: An update on the dynamics of pathogen survival/transport as well as practical risk mitigation strategies. *Int. J. Hyg. Environ. Health* 227, 113524.
- Alexis, J., de, Kerchove,.., Elimelech, M., 2007. Formation of Polysaccharide Gel Layers in the Presence of  $\text{Ca}^{2+}$  and  $\text{K}^+$  Ions: Measurements and Mechanisms. *Macromolecules* 8 (1), 113–121.
- Alizadeh-Pasdar, N., Li-Chan, E.C.Y., 2000. Comparison of Protein Surface Hydrophobicity Measured at Various pH Values Using Three Different Fluorescent Probes. *J. Agric. Food Chem.* 48 (2), 328–334.
- Allen, A.S., Borchardt, M.A., Kieke, B.A., Dunfield, K.E., Parker, B.L., 2017. Virus occurrence in private and public wells in a fractured dolostone aquifer in Canada. *Hydrogeol. J.* 25 (4), 1117–1136.
- Aller, L., Bennett, T., Lehr, J.H., Petty, R.J., Hackett, G., 1987. DRASTIC : A standardized system for evaluating groundwater pollution potential using hydrogeologic settings. *J. Geol. Soc. India* 29 (1).
- Anders, R., Chrysikopoulos, C.V., 2006. Evaluation of the factors controlling the time-dependent inactivation rate coefficients of bacteriophage MS2 and PRD1. *Environ. Sci. Technol.* 40 (10), 3237–3242.
- Ando, Y., Tsuzuki, T., 1984. The role of surface charge in ionic germination of *Clostridium perfringens* spores. *J. Gen. Microbiol.* 130 (2), 267–273.
- Arcega-Cabrera, F., Sickman, J.O., Fargher, L., Herrera-Silveira, J., Lucero, D., Oceguera-Vargas, I., Lamas-Cosio, E., Robledo-Ardila, P.A., 2021. Groundwater Quality in the Yucatan Peninsula: Insights from Stable Isotope and Metals Analysis. *Groundwater* 59, 878–891.
- Armanious, A., Aepli, M., Jacak, R., Refardt, D., Sigstam, T., Kohn, T., Sander, M., 2016. Viruses at Solid-Water Interfaces: A Systematic Assessment of Interactions Driving Adsorption. *Environ. Sci. Technol.* 50 (2), 732–743.
- Aronino, R., Dragy, C., Arkhangelsky, E., Shandalov, S., Oron, G., Brenner, A., Gitit, V., 2009. Removal of viruses from surface water and secondary effluents by sand filtration. *Water Res.* 43 (1), 87–96.
- Babich, H., Stotzky, G., 1980. Reductions in inactivation rates of bacteriophages by clay minerals in lake water. *Water Res.* 14 (2), 185–187.
- Bai, H., Cochet, N., Pauss, A., Lamy, E., 2016. Bacteria cell properties and grain size impact on bacteria transport and deposition in porous media. *Colloids Surf. B139*, 148–155.
- Baker, A.T., Davies, J.A., Bates, E.A., Moses, E., Mundy, R.M., Marlow, G., Cole, D.K., Bliss, C.M., Rizkallah, P.J., Parker, A.L., 2021. The Fiber Knob Protein of Human Adenovirus Type 49 Mediates Highly Efficient and Promiscuous Infection of Cancer Cell Lines Using a Novel Cell Entry Mechanism. *J. Virol.* 95 (4), e01849–01820.
- Balboa, S., Mauricio-Iglesias, M., Rodriguez, S., Martínez-Lamas, L., Vasallo, F.J., Regueiro, B., Lema, J.M., 2021. The fate of SARS-CoV-2 in WWTPS points out the sludge line as a suitable spot for detection of COVID-19. *Sci. Total Environ.* 772, 145268.
- Bales, R.C., Hinkle, S.R., Kroeger, T.W., Stocking, K., Gerba, C.P., 1991. Bacteriophage adsorption during transport through porous media: chemical perturbations and reversibility. *Environ. Sci. Technol.* 25 (12), 2088–2095.

- Balkwill, D.L., Ghiorse, W.C., 1985. Characterization of subsurface bacteria associated with two shallow aquifers in Oklahoma. *Appl. Environ. Microbiol.* 50 (3), 580–588.
- Baltus, R.E., Badireddy, A.R., Delavari, A., Chellam, S., 2017. Free Diffusivity of Icosahedral and Tailed Bacteriophages: Experiments, Modeling, and Implications for Virus Behavior in Media Filtration and Flocculation. *Environ. Sci. Technol.* 51 (3), 1433–1440.
- Barbulescu, A., 2020. Assessing Groundwater Vulnerability: DRASTIC and DRASTIC-Like Methods: A Review. *Water* 12 (5), 1356.
- Baygents, J.C., Glynn, J.R., Albinger, O., Biesemeyer, B.K., Ogden, K.L., Arnold, R.G., 1998. Variation of Surface Charge Density in Monoclonal Bacterial Populations: Implications for Transport through Porous Media. *Environ. Sci. Technol.* 32 (11), 1596–1603.
- Bellamy, W.D., Hendricks, D.W., Logsdon, G.S., 1985. Slow Sand Filtration: Influences of Selected Process Variables. *J. - Am. Water Works Assoc.* 77 (12), 62–66.
- Bellou, M.I., Syrigouna, V.I., Tselepi, M.A., Koklino, P.A., Paparodopoulos, S.C., Vantarakis, A., Chrysikopoulos, C.V., 2015. Interaction of human adenoviruses and coliphages with kaolinite and bentonite. *Sci. Total Environ.* 517, 86–95.
- Bidhuri, S., Taqi, M., Khan, M.M.A., 2018. Water-borne disease: Link between human health and water use in the Mithapur and Jaitpur area of the NCT of Delhi. *Journal of Public Health-Heidelberg* 26 (1), 119–126.
- Blanco, A., Abid, I., Al-Otaibi, N., Perez-Rodriguez, F.J., Fuentes, C., Guix, S., Pinto, R.M., Bosch, A., 2019. Glass Wool Concentration Optimization for the Detection of Enveloped and Non-enveloped Waterborne Viruses. *Food Environ. Virol.* 11 (2), 184–192.
- Borchardt, M.A., Bradbury, K.R., Alexander Jr., E.C., Kolberg, R.J., Alexander, S.C., Archer, J.R., Braatz, L.A., Forest, B.M., Green, J.A., Spencer, S.K., 2011. Norovirus Outbreak Caused by a New Septic System in a Dolomite Aquifer. *Groundwater* 49 (1), 85–97.
- Bradford, S.A., Morales, V.L., Zhang, W., Harvey, R.W., Packman, A.I., Mohanram, A., Welty, C., 2013. Transport and Fate of Microbial Pathogens in Agricultural Settings. *Crit. Rev. Environ. Sci. Technol.* 43 (8), 775–893.
- Bradford, S.A., Simunek, J., Walker, S.L., 2006a. Transport and straining of *E. coli* O157:H7 in saturated porous media. *Water Resour. Res.* 42 (12), W12S12.
- Bradford, S.A., Tadassa, Y.F., Jin, Y., 2006b. Transport of Coliphage in the Presence and Absence of Manure Suspension. *J. Environ. Qual.* 35 (5), 1692–1701.
- Bradford, S.A., Torkzaban, S., Wiegmann, A., 2011. Pore-Scale Simulations to Determine the Applied Hydrodynamic Torque and Colloid Immobilization. *Vadose Zone J.* 10 (1), 252–261.
- Buckerfield, S.J., Quilliam, R.S., Bussiere, L., Waldron, S., Naylor, L.A., Li, S., Oliver, D.M., 2020. Chronic urban hotspots and agricultural drainage drive microbial pollution of karst water resources in rural developing regions. *Sci. Total Environ.* 744, 140898.
- Buckerfield, S.J., Quilliam, R.S., Waldron, S., Naylor, L.A., Li, S., Oliver, D.M., 2019. Rainfall-driven *E. coli* transfer to the stream-conduit network observed through increasing spatial scales in mixed land-use paddy farming karst terrain. *Water Res.* X 5, 100038.
- Cai, P., Huang, Q., Walker, S.L., 2013. Deposition and Survival of *Escherichia coli* O157:H7 on Clay Minerals in a Parallel Plate Flow System. *Environ. Sci. Technol.* 47 (4), 1896–1903.
- Cao, H., Tsai, F.T.C., Rusch, K.A., 2010. Salinity and Soluble Organic Matter on Virus Sorption in Sand and Soil Columns. *Groundwater* 48 (1), 42–52.
- Carrascosa, J.L., Kellenberger, E., 1978. Head maturation pathway of bacteriophages T4 and T2. III. Isolation and characterization of particles produced by mutants in gene 17. *J. Virol.* 25 (3), 831–844.
- Castro, F.D., Tufenkji, N., 2007. Relevance of Nontoxicogenic Strains as Surrogates for *Escherichia coli* O157:H7 in Groundwater Contamination Potential: Role of Temperature and Cell Acclimation Time. *Environ. Sci. Technol.* 41 (12), 4332–4338.
- Chen, G., Walker, S.L., 2007. Role of Solution Chemistry and Ion Valence on the Adhesion Kinetics of Groundwater and Marine Bacteria. *Langmuir* 23 (13), 7162–7169.
- Cheng, H., Hu, E., Hu, Y., 2012. Impact of mineral micropores on transport and fate of organic contaminants: A review. *J. Contam. Hydrol.* 129–130, 80–90.
- Cheng, P.K.C., Wong, D.K.K., Chung, T.W.H., Lim, W.W.L., 2005. Norovirus contamination found in oysters worldwide. *J. Med. Virol.* 76 (4), 593–597.
- Chrysikopoulos, C.V., Aravantinou, A.F., 2014. Virus attachment onto quartz sand: Role of grain size and temperature. *J. Environ. Chem. Eng.* 2 (2), 796–801.
- Chrysikopoulos, C.V., Syrigouna, V.I., 2012. Attachment of bacteriophages MS2 and ΦX174 onto kaolinite and montmorillonite: extended-DLVO interactions. *Colloids Surf B Biointerfaces* 92, 74–83.
- Chrysikopoulos, C.V., Masciopinto, C., La Mantia, R., Manariotis, I.D., 2010. Removal of Biocolloids Suspended in Reclaimed Wastewater by Injection into a Fractured Aquifer Model. *Environ. Sci. Technol.* 44 (3), 971–977.
- Chu, Y., Jin, Y., Flury, M., Yates, M.V., 2001. Mechanisms of virus removal during transport in unsaturated porous media. *Water Resour. Res.* 37 (2), 253–263.
- Collins, K.E., Cronin, A.A., Rueedi, J., Pedley, S., Joyce, E., Humble, P.J., Tellam, J.H., 2006. Fate and transport of bacteriophage in UK aquifers as surrogates for pathogenic viruses. *Eng. Geol.* 85 (1–2), 33–38.
- Corapcioglu, M., Vogel, J., Munster, C., Pillai, S., Dowd, S., Wang, S., 2006. Virus transport experiments in a sandy aquifer. *Water Air Soil Pollut.* 169 (1–4), 47–65.
- Cronin, A.A., Taylor, R.G., Powell, K.L., Barrett, M.H., Trowsdale, S.A., Lerner, D.N., 2003. Temporal variations in the depth-specific hydrochemistry and sewage-related microbiology of an urban sandstone aquifer, Nottingham, United Kingdom. *Hydrogeol. J.* 11 (2), 205–216.
- Cui, L., Chen, P.Y., Zhang, B.F., Zhang, D.Y., Li, J.Y., Martin, F.L., Zhang, K.S., 2015. Interrogating chemical variation via layer-by-layer SERS during biofouling and cleaning of nanofiltration membranes with further investigations into cleaning efficiency. *Water Res.* 87, 282–291.
- Dang, H.T.T., Tarabara, V.V., 2021. Attachment of human adenovirus onto household paints. *Colloids Surf. B* 204, 111812.
- da Silva, A.K., Kavanagh, O.V., Estes, M.K., Elimelech, M., 2011. Adsorption and Aggregation Properties of Norovirus GI and GII Virus-like Particles Demonstrate Differing Responses to Solution Chemistry. *Environ. Sci. Technol.* 45 (2), 520–526.
- Deblois, R.W., Wesley, R.K., 1977. Sizes and concentrations of several type C noroviruses and bacteriophage T2 by the resistive-pulse technique. *J. Virol.* 23 (2), 227–233.
- Degréve, L., Fuzo, C.A., Caliri, A., 2012. Extensive structural change of the envelope protein of dengue virus induced by a tuned ionic strength: conformational and energetic analyses. *J. Comput. Aided Mol. Des.* 26 (12), 1311–1325.
- Dieter, C.A., Maupin, M.A., 2017. Public Supply and Domestic Water Use in the United States, 2015, No. 2017-1131. US Geological Survey.
- Dika, C., Duval, J.F.L., Francius, G., Perrin, A., Gantzer, C., 2015. Isoelectric point is an inadequate descriptor of MS2, Phi X 174 and PRD1 phages adhesion on abiotic surfaces. *J. Colloid Interface Sci.* 446, 327–334.
- Dika, C., Ly-Chatain, M.H., Francius, G., Duval, J.F.L., Gantzer, C., 2013. Non-DLVO adhesion of F-specific RNA bacteriophages to abiotic surfaces: Importance of surface roughness, hydrophobic and electrostatic interactions. *Colloids Surf. A* 435, 178–187.
- Dion, M.B., Oechslin, F., Moineau, S., 2020. Phage diversity, genomics and phylogeny. *Nat. Rev. Microbiol.* 18 (3), 125–138.
- Dong, Z., Yang, H.Y., Wu, D., Ni, J.R., Kim, H., Tong, M.P., 2014. Influence of silicate on the transport of bacteria in quartz sand and iron mineral-coated sand. *Colloids Surf. B* 123, 995–1002.
- Dowd, S.E., Pillai, S.D., 1997. Survival and transport of selected bacterial pathogens and indicator viruses under sandy aquifer conditions. *J. Environ. Sci. Health, Part A* 32 (8), 2245–2258.
- Dowd, S.E., Pillai, S.D., Wang, S., Corapcioglu, M.Y., 1998. Delineating the specific influence of virus isoelectric point and size on virus adsorption and transport through sandy soils. *Appl. Environ. Microbiol.* 64 (2), 405–410.
- Du, B., Gu, Y., Chen, G., Wang, G., Liu, L., 2020. Flagellar motility mediates early-stage biofilm formation in oligotrophic aquatic environment. *Ecotoxicol. Environ. Saf.* 194, 110340.
- Farkas, K., Varsani, A., Pang, L., 2015. Adsorption of Rotavirus, MS2 Bacteriophage and Surface-Modified Silica Nanoparticles to Hydrophobic Matter. *Food Environ. Virol.* 7 (3), 261–268.
- Feldner, J., Bredt, W., Kahane, I., 1983. Influence of cell shape and surface charge on attachment of *Mycoplasma pneumoniae* to glass surfaces. *J. Bacteriol.* 153 (1), 1–5.
- Feng, H.R., Ruan, Y.F., Wu, R.B., Zhang, H.Y., Lam, P.K.S., 2019. Occurrence of disinfection by-products in sewage treatment plants and the marine environment in Hong Kong. *Ecotoxicol. Environ. Saf.* 181, 404–411.
- Fernando, G., Hernández, P., Burgos, J., Sanz, B., Ordóñez, J., 1991. Extracellular proteinase from *Enterococcus faecalis* subsp. *liquefaciens*. II. Partial purification and some technological important properties. *Folia Microbiologica* 36 (5), 429–436.
- Flint, S.J., Enquist, L., Racaniello, V., Skalka, A., 2004. Principles of virology: molecular biology, pathogenesis, and control of animal viruses, Second Edition. AMS Press.
- Flury, M., Aramrak, S., 2017. Role of air-water interfaces in colloid transport in porous media: A review. *Water Resour. Res.* 53 (7), 5247–5275.
- Fong, T.-T., Mansfield, L.S., Wilson, D.L., Schwab, D.J., Molloy, S.L., Rose, J.B., 2007. Massive microbiological groundwater contamination associated with a waterborne outbreak in Lake Erie, South Bass Island, Ohio. *Environ. Health Perspect.* 115 (6), 856–864.
- Foppen, J.W., Liem, Y., Schijven, J., 2008. Effect of humic acid on the attachment of *Escherichia coli* in columns of goethite-coated sand. *Water Res.* 42 (1), 211–219.
- Foppen, J.W.A., Okletsey, S., Schijven, J.F., 2006. Effect of goethite coating and humic acid on the transport of bacteriophage PRD1 in columns of saturated sand. *J. Contam. Hydrol.* 85 (3), 287–301.
- Foster, S.S.D., Chilton, P.J., 2004. Downstream of downtown: urban wastewater as groundwater recharge. *Hydrogeol. J.* 12 (1), 115–120.
- Fout, G.S., Borchardt, M.A., Kieke, B.A., Karim, M.R., 2017. Human virus and microbial indicator occurrence in public-supply groundwater systems: meta-analysis of 12 international studies. *Hydrogeol. J.* 25 (4), 903–919.
- Fouz, N., Pangestri, K.N.A., Yasir, M., Al-Malki, A.L., Azhar, E.I., Hill-Cawthorne, G.A., Abd El Ghany, M., 2020. The Contribution of Wastewater to the Transmission of Antimicrobial Resistance in the Environment: Implications of Mass Gathering Settings. *Trop Med Infect Dis.* 5 (1), 33.
- Fumian, T.M., Leite, J.P.G., Castello, A.A., Gaggero, A., Caillou, M.S.L.D., Miagostovich, M.P., 2010. Detection of rotavirus A in sewage samples using multiplex qPCR and an evaluation of the ultracentrifugation and adsorption-elution methods for virus concentration. *J. Virol. Methods* 170 (1), 42–46.
- Gallardo-Moreno, A.M., González-Martín, M.L., Bruque, J.M., Pérez-Giraldo, C., Sánchez-Silos, R., Gómez-García, A.C., 2003. Influence of the growth medium, suspending liquid and measurement temperature on the physico-chemical surface properties of two enterococci strains. *J. Adhes. Sci. Technol.* 17 (14), 1877–1887.

- Gargiulo, G., Bradford, S.A., Simunek, J., Ustohal, P., Vereecken, H., Klumpp, E., 2008. Bacteria Transport and Deposition under Unsaturated Flow Conditions: The Role of Water Content and Bacteria Surface Hydrophobicity. *Vadose Zone J.* 7 (2), 406–419.
- Gasteiger, E., Gattiker, A., Hoogland, C., Ivanyi, I., Appel, R.D., Bairoch, A., 2003. ExPASy: the proteomics server for in-depth protein knowledge and analysis. *Nucleic Acids Res.* 31 (13), 3784–3788.
- Gerba, C.P., Betancourt, W.Q., Kitajima, M., 2017. How much reduction of virus is needed for recycled water: A continuous changing need for assessment? *Water Res.* 108, 25–31.
- Ghanem, N., Kiesel, B., Kallies, R., Harms, H., Chatzinotas, A., Wick, L.Y., 2016. Marine Phages As Tracers: Effects of Size, Morphology, and Physico-Chemical Surface Properties on Transport in a Porous Medium. *Environ. Sci. Technol.* 50 (23), 12816–12824.
- Gharabaghi, B., Safadoust, A., Mahboubi, A.A., Mosaddeghi, M.R., Unc, A., Ahrens, B., Sayyad, G., 2015. Temperature effect on the transport of bromide and *E. coli* NAR in saturated soils. *J. Hydrol.* 522, 418–427.
- Gilbert, P., Evans, D.J., Evans, E., Duguid, I.G., Brown, M.R., 1991. Surface characteristics and adhesion of *Escherichia coli* and *Staphylococcus epidermidis*. *J. Appl. Bacteriol.* 71 (1), 72–77.
- Gitis, V., Adin, A., Nasser, A., Gun, J., Lev, O., 2002. Fluorescent dye labeled bacteriophages—a new tracer for the investigation of viral transport in porous media: 1. Introduction and characterization. *Water Research* 36 (17), 4227–4234.
- Gomez-Suarez, C., Busscher, H.J., vander, Mei, H.C., 2001. Analysis of Bacterial Detachment from Substratum Surfaces by the Passage of Air-Liquid Interfaces. *Appl. Environ. Microbiol.* 67 (6), 2531–2537.
- Gordesli, F.P., Abu-Lail, N.I., 2012. Combined Poisson and soft-particle DLVO analysis of the specific and nonspecific adhesion forces measured between *L. monocytogenes* grown at various temperatures and silicon nitride. *Environ. Sci. Technol.* 46 (18), 10089–10098.
- Graciaa, A., Morel, G., Saulner, P., Lachaise, J., Schechter, R., 1995. The  $\zeta$ -potential of gas bubbles. *J. Colloid Interface Sci.* 172 (1), 131–136.
- Gunawardhana, L., So, K., Masaki, S., 2009. Seasonal Change of Groundwater Flow and its Effect on Temperature Distribution in Sendai Plain. *Advances in Water Resources and Hydraulic Engineering* 1 (6), 193–198.
- Gundy, P.M., Gerba, C.P., Pepper, I.L., 2009. Survival of Coronaviruses in Water and Wastewater. *Food Environ. Virol.* 1 (1), 10–14.
- Han, J., Jin, Y., Wilson, C.S., 2006. Virus retention and transport in chemically heterogeneous porous media under saturated and unsaturated flow conditions. *Environ. Sci. Technol.* 40 (5), 1547–1555.
- Haramoto, E., Katayama, H., Oguma, K., Ohgaki, S., 2005. Application of cation-coated filter method to detection of noroviruses, enteroviruses, adenoviruses, and torque teno viruses in the Tamagawa River in Japan. *Appl. Environ. Microbiol.* 71 (5), 2403–2411.
- Haveman, S., 2002. Distribution of culturable microorganisms in Fennoscandian Shield groundwater. *FEMS Microbiol. Ecol.* 39 (2), 129–137.
- Haznedaroglu, B.Z., Zorlu, O., Hill, J.E., Walker, S.L., 2010. Identifying the Role of Flagella in the Transport of Motile and Nonmotile *Salmonella enterica* Serovars. *Environ. Sci. Technol.* 44 (11), 4184–4190.
- He, Q., Wu, Q.-Q., Ma, H.-F., Zhou, Z.-M., Yuan, B.-L., 2014. Effects of algae and kaolinite particles on the survival of bacteriophage MS2. *Environ. Sci.* 35 (8), 3192–3197.
- Hellmér, M., Paxéus, N., Magnus, L., Enache, L., Arnholm, B., Johansson, A., Bergström, T., Norder, H., 2014. Detection of Pathogenic Viruses in Sewage Provided Early Warnings of Hepatitis A Virus and Norovirus Outbreaks. *Appl. Environ. Microbiol.* 80 (21), 6771–6781.
- Henderson, I., 1997. Antigen 43, a phase-variable bipartite outer membrane protein, determines colony morphology and autoaggregation in *Escherichia coli* K-12. *149* (1), 115–120.
- Hynds, P.D., Thomas, M.K., Pintar, K.D.M., 2014. Contamination of Groundwater Systems in the US and Canada by Enteric Pathogens, 1990–2013: A Review and Pooled-Analysis. *PLoS ONE* 9, (5) e93301.
- Jacobs, A., Lafolie, F., Henry, J.M., Debroux, M., 2007. Kinetic adhesion of bacterial cells to sand: Cell surface properties and adhesion rate. *Colloids Surf. B* 59 (1), 35–45.
- Jiang, S.C., Bai, B., 2018. Influence of particle shape on the suspended particle transport and deposition in porous media. *Rock Soil Mech.* 39 (6), 2043–2051.
- Jimenez-Zaragoza, M., Yubero, M.P.L., Martin-Forero, E., Caston, J.R., Reguera, D., Luque, D., de Pablo, P.J., Rodriguez, J.M., 2018. Biophysical properties of single rotavirus particles account for the functions of protein shells in a multilayered virus. *eLife* 7, e37295.
- Jin, Y., Flury, M., 2002. Fate and transport of viruses in porous media. *Adv. Agron.* 77, 39–102.
- Jin, Y., Pratt, E., Yates, M.V., 2000. Effect of Mineral Colloids on Virus Transport through Saturated Sand Columns. *J. Environ. Qual.* 29 (2), 532–539.
- Kamezawa, Y., Nakahara, T., Nakano, S., Abe, Y., Isono, T., 1997. Streptococcal mitogenic exotoxin Z, a novel acidic superantigenic toxin produced by a T1 strain of *Streptococcus pyogenes*. *Infect. Immun.* 65 (9), 3828–3833.
- Kampf, G., Todt, D., Pfaender, S., Steinmann, E., 2020. Persistence of coronaviruses on inanimate surfaces and their inactivation with biocidal agents. *J. Hosp. Infect.* 104 (3), 246–251.
- Kamrani, S., Rezaei, M., Kord, M., Baalousha, M., 2018. Transport and retention of carbon dots (CDs) in saturated and unsaturated porous media: Role of ionic strength, pH, and collector grain size. *Water Res.* 133, 338–347.
- Kang, Z.-H., Wang, L., Zheng, G.-H., Zhou, Q., 2006. Capability of  $Cu^{2+}$  adsorption on cell surface of *Pseudomonas putida* 5-x. *Environ. Sci.* 27 (5), 965–971.
- Kar, S., Maity, J.P., Jean, J.-S., Liu, C.-C., Nath, B., Yang, H.-J., Bundschuh, J., 2010. Arsenic-enriched aquifers: Occurrences and mobilization of arsenic in groundwater of Ganges Delta Plain, Barasat, West Bengal, India. *Appl. Geochem.* 25 (12), 1805–1814.
- Katzourakis, V.E., Chrysikopoulos, C.V., 2014. Mathematical modeling of colloid and virus cotransport in porous media: Application to experimental data. *Adv. Water Resour.* 68, 62–73.
- Ke, Z.L., Oton, J.Q., Qu, K., Cortese, M., Zila, V., McKeane, L., Nakane, T., Zivanov, J., Neufeldt, C.J., Cerikan, B., Lu, J.M., Peukes, J., Xiong, X.L., Krausslich, H.G., Scheres, S.H.W., Bartenschlager, R., Briggs, J.A.G., 2020. Structures and distributions of SARS-CoV-2 spike proteins on intact virions. *Nature* 588 (7838), 498–502.
- Kim, H.N., Bradford, S.A., Walker, S.L., 2009. *Escherichia coli* O157:H7 Transport in Saturated Porous Media: Role of Solution Chemistry and Surface Macromolecules. *Environ. Sci. Technol.* 43 (12), 4340–4347.
- Kim, H.N., Walker, S.L., 2009. *Escherichia coli* transport in porous media: Influence of cell strain, solution chemistry, and temperature. *Colloids Surf. B* 71 (1), 160–167.
- Kim, M.-K., Kim, S.-B., Park, S.-J., 2008. Bacteria transport in an unsaturated porous media: incorporation of air-water interface area model into transport modelling. *Hydrol. Process.* 22 (13), 2370–2376.
- Klemm, P., Hjerrild, L., Gjermansen, M., Schembri, M.A., 2003. Structure-function analysis of the self-recognizing Antigen 43 autotransporter protein from *Escherichia coli*. *Mol. Microbiol.* 51 (1), 283–296.
- Ko, C.H., Elimelech, M., 2000. The 'shadow effect' in colloid transport and deposition dynamics in granular porous media: Measurements and mechanisms. *Environ. Sci. Technol.* 34 (17), 3681–3689.
- Kokkinos, P., Syrigouna, V.I., Tslepi, M.A., Bellou, M., Chrysikopoulos, C.V., Vantarakis, A., 2015. Transport of Human Adenoviruses in Water Saturated Laboratory Columns. *Food Environ. Virol.* 7 (2), 122–131.
- Kortelainen, N.M., Karhu, J.A., 2009. Geochemical and isotopic evolution of high-pH groundwater in a carbonate-bearing glacigenic aquifer. SW Finland. *Hydrol. Res.* 40 (1), 19–31.
- Krishna, Darbha, G., Fischer, C., Michler, A., Luetzenkirchen, J., Schafer, T., Heberling, F., Schild, D., 2012. Deposition of latex colloids at rough mineral surfaces: An analogue study using nanopatterned surfaces. *Langmuir* 28 (16), 6606–6617.
- Kumar, M., Thakur, A.K., Mazumder, P., Kuroda, K., Mohapatra, S., Rinklebe, J., Ramanathan, A., Cetecioglu, Z., Jain, S., Tyagi, V.K., Gikas, P., Chakraborty, S., Tahmidul Islam, M., Ahmad, A., Shah, A.V., Patel, A.K., Watanabe, T., Vithanage, M., Bibby, K., Kitajima, M., Bhattacharya, P., 2020. Frontier review on the propensity and repercussion of SARS-CoV-2 migration to aquatic environment. *J. Hazard. Mater. Lett.* 1, 100001.
- Kvitsand, H.M.L., Fiksdal, L., 2010. Waterborne disease in Norway: emphasizing outbreaks in groundwater systems. *Water Sci. Technol.* 61 (3), 563–571.
- La Rosa, G., Iaconelli, M., Mancini, P., Bonanno Ferraro, G., Veneri, C., Bonadonna, L., Lucentini, L., Suffredini, E., 2020. First detection of SARS-CoV-2 in untreated wastewaters in Italy. *Sci. Total Environ.* 736, 139652.
- Langlet, J., Gaboriaud, F., Gantzer, C., Duval, J.F.L., 2008. Impact of chemical and structural anisotropy on the electrophoretic mobility of spherical soft multilayer particles: The case of bacteriophage MS2. *Biophys. J.* 94 (8), 3293–3312.
- Lee, H., Kim, M., Lee, J.E., Lim, M., Kim, M., Kim, J.-M., Jheong, W.-H., Kim, J., Ko, G., 2011. Investigation of norovirus occurrence in groundwater in metropolitan Seoul. *Korean Sci. Total Environ.* 409 (11), 2078–2084.
- Li, P.Y., Li, L., Yang, K.X., Zheng, T.L., Liu, J.X., Wang, Y.J., 2021. Characteristics of microbial aerosol particles dispersed downwind from rural sanitation facilities: Size distribution, source tracking and exposure risk. *Environ. Res.* 195, 110798.
- Li, S., 2018. Study on environmental behavior characteristics of viruses in vadose zone, Jilin University, Master Thesis. (in Chinese with English abstract).
- Li, T.T., Shen, C.Y., Wu, S., Jin, C., Bradford, S.A., 2020. Synergies of surface roughness and hydration on colloid detachment in saturated porous media: Column and atomic force microscopy studies. *Water Res.* 183, 116068.
- Liu, G.S., Fang, Z.D., Zhong, H., Shi, L.S., Yang, X., Liu, Z.F., 2020. Transport of *Pseudomonas aeruginosa* in Porous Media Mediated by Low-Concentration Surfactants: The Critical Role of Surfactant to Change Cell Surface Hydrophobicity. *Water Resour. Res.* 56, (2), e2019WR026103.
- Liu, L.L., Hu, J.Y., Hou, Y.X., Tao, Z., Chen, Z.H., Chen, K., 2021a. Pit latrines may be a potential risk in rural China and low-income countries when dealing with COVID-19. *Sci. Total Environ.* 761, 143283.
- Liu, Q., Lazouskaya, V., He, Q.X., Jin, Y., 2010. Effect of Particle Shape on Colloid Retention and Release in Saturated Porous Media. *J. Environ. Qual.* 39 (2), 500–508.
- Liu, Y., Yang, C.-H., Li, J., 2007. Influence of Extracellular Polymeric Substances on *Pseudomonas aeruginosa* Transport and Deposition Profiles in Porous Media. *Environ. Sci. Technol.* 41 (1), 198–205.
- Liu, Y.N., Lv, Z.T., Yang, S.Y., Liu, X.W., 2021b. Optical Tracking of the Interfacial Dynamics of Single SARS-CoV-2 Pseudoviruses. *Environ. Sci. Technol.* 55 (7), 4115–4122.
- Liu Ting, X.Q., 2012. Influence of alkaline environment on the hydrophobicity of starved *Enterococcus faecalis*. *J. South. Med. Univ.* 32 (4), 681–684.
- Luo, X., Xiang, X.Y., Yang, Y.H., Huang, G.Y., Fu, K.D., Che, R.X., Chen, L.Q., 2020. Seasonal effects of river flow on microbial community coalescence and diversity in a riverine network. *FEMS Microbiol. Ecol.* 96 (8), fiaa132.
- Lutterodt, G., Basnet, M., Foppen, J.W.A., Uhlenbrook, S., 2009. The effect of surface characteristics on the transport of multiple *Escherichia coli* isolates in large scale columns of quartz sand. *Water Res.* 43 (3), 595–604.

- Lv, X.Y., Gao, B., Sun, Y.Y., Dong, S.N., Wu, J.C., Jiang, B.L., Shi, X.Q., 2016. Effects of grain size and structural heterogeneity on the transport and retention of nano-TiO<sub>2</sub> in saturated porous media. *Sci. Total Environ.* 563, 987–995.
- Lytle, D.A., Rice, E.W., Johnson, C.H., Fox, K.R., 1999. Electrophoretic Mobilities of *Escherichia coli* O157:H7 and Wild-Type *Escherichia coli* Strains. *Appl. Environ. Microbiol.* 65 (7), 3222–3225.
- Ma, H., Bolster, C., Johnson, W.P., Li, K., Pazmino, E., Camacho, K.M., Anselmo, A.C., Mitragotri, S., 2020. Coupled influences of particle shape, surface property and flow hydrodynamics on rod-shaped colloid transport in porous media. *J. Colloid Interface Sci.* 577, 471–480.
- Makowska, N., Breska, K., Koczura, R., Philips, A., Nowis, K., Mokracka, J., 2021. Urban wastewater as a conduit for pathogenic Gram-positive bacteria and genes encoding resistance to beta-lactams and glycopeptides. *Sci. Total Environ.* 765, 144176.
- Marinova, K., Alargova, R., Denkov, N., Velev, O., Petsev, D., Ivanov, I., Borwankar, R., 1996. Charging of oil-water interfaces due to spontaneous adsorption of hydroxyl ions. *Langmuir* 12 (8), 2045–2051.
- McCarthy, J.F., McKay, L.D., 2004. Colloid transport in the subsurface: Past, present, and future challenges. *Vadose Zone J.* 3 (2), 326–337.
- McClaine, J.W., Ford, R.M., 2002. Characterizing the adhesion of motile and nonmotile *Escherichia coli* to a glass surface using a parallel-plate flow chamber. *Biotechnol. Bioeng.* 78 (2), 179–189.
- Meinders, J.M., van der Mei, H.C., Busscher, H.J., 1994. Physicochemical Aspects of Deposition of *Streptococcus thermophilus* B to Hydrophobic and Hydrophilic Substrata in a Parallel Plate Flow Chamber. *J. Colloid Interface Sci.* 164 (2), 355–363.
- Miao, R., Wang, L., Mi, N., Gao, Z., Liu, T., Lv, Y., Wang, X., Meng, X., Yang, Y., 2015. Enhancement and Mitigation Mechanisms of Protein Fouling of Ultrafiltration Membranes under Different Ionic Strengths. *Environ. Sci. Technol.* 49 (11), 6574–6580.
- Millsap, K., Reid, G., van der Mei, H.C., Busscher, H.J., 1994. Displacement of *Enterococcus faecalis* from hydrophobic and hydrophilic substrata by *Lactobacillus* and *Streptococcus* spp. as studied in a parallel plate flow chamber. *Appl. Environ. Microbiol.* 60 (6), 1867–1874.
- Mohanty, S.K., Boehm, A.B., 2014. Escherichia coli Removal in Biochar-Augmented Biofilter: Effect of Infiltration Rate, Initial Bacterial Concentration, Biochar Particle Size, and Presence of Compost. *Environ. Sci. Technol.* 48 (19), 11535–11542.
- Mohanty, S.K., Cantrell, K.B., Nelson, K.L., Boehm, A.B., 2014. Efficacy of biochar to remove Escherichia coli from stormwater under steady and intermittent flow. *Water Res.* 61, 288–296.
- Morales, I., Amador, J.A., Boving, T., 2015. Bacteria Transport in a Soil-Based Wastewater Treatment System under Simulated Operational and Climate Change Conditions. *J. Environ. Qual.* 44 (5), 1459–1472.
- Morales, V.L., Gao, B., Steenhuis, T.S., 2009. Grain Surface-Roughness Effects on Colloidal Retention in the Vadose Zone. *Vadose Zone J.* 8 (1), 11–20.
- Morse, S.S., Mazet, J.A.K., Woolhouse, M., Parrish, C.R., Carroll, D., Karesh, W.B., Zambrana-Torrello, C., Lipkin, W.I., Daszak, P., 2012. Zoonoses 3 Prediction and prevention of the next pandemic zoonosis. *Lancet* 380 (9857), 1956–1965.
- Murphy, F.A., 1998. Emerging zoonoses. *Emerg. Infect. Dis.* 4 (3), 429–435.
- Murphy, H.M., Prioleau, M.D., Borchardt, M.A., Hynds, P.D., 2017. Review: Epidemiological evidence of groundwater contribution to global enteric disease, 1948–2015. *Hydrogeol. J.* 25 (4), 981–1001.
- Nikaido, H., 2003. Molecular basis of bacterial outer membrane permeability revisited. *Microbiol. Mol. Biol. Rev.* 67 (4), 593–656.
- Nola, M., Ewoti, O.V.N., Nougang, M., Moungang, M.L., Chihib, N.E., Krier, F., Servais, P., Hornez, J.P., Njine, T., 2010. Involvement of cell shape and flagella in the bacterial retention during percolation of contaminated water through soil columns in tropical region. *J. Environ. Sci. Health Part A-Toxic/Hazard. Subst. Environ. Eng.* 45 (11), 1297–1306.
- Oh, Y.J., Jo, W., Yang, Y., Park, S., 2007. Influence of culture conditions on Escherichia coli O157:H7 biofilm formation by atomic force microscopy. *Ultramicroscopy* 107 (10–11), 869–874.
- Oliveira, H., Santos, C., Paterson, R., Gusmão, N., Lima, N., 2016. Fungi from a groundwater-fed drinking water supply system in Brazil. *Int. J. Env. Res. Public Health* 13 (3), 304.
- Ong, S.W.X., Tan, Y.K., Chia, P.Y., Lee, T.H., Ng, O.T., Wong, M.S.Y., Marimuthu, K., 2020. Air, Surface Environmental, and Personal Protective Equipment Contamination by Severe Acute Respiratory Syndrome Coronavirus 2 (SARS-CoV-2) From a Symptomatic Patient. *J. Am. Med. Assoc.* 323 (16), 1610–1612.
- Panda, K.P., Upadhyay, A., Jha, M.K., Sharma, S.P., 2020. Mapping of laterite zones using 2D electrical resistivity tomography survey in parts of Paschim Medinipur, West Bengal, India: An approach for artificial groundwater recharge. *J. Earth Syst. Sci.* 129 (1), 119.
- Pandey, L.M., 2020. Surface engineering of personal protective equipments (PPEs) to prevent the contagious infections of SARS-CoV-2. *Surf. Eng.* 36 (9), 901–907.
- Pandey, P.K., Kass, P.H., Soupir, M.L., Biswas, S., Singh, V.P., 2014. Contamination of water resources by pathogenic bacteria. *AMB Express* 4 (1), 51.
- Pang, L., Farkas, K., Lin, S., Hewitt, J., Premaratne, A., Close, M., 2021. Attenuation and transport of human enteric viruses and bacteriophage MS2 in alluvial sand and gravel aquifer media—laboratory studies. *Water Res.* 196, 117051.
- Pang, L., Nowostawska, U., Ryan, J.N., Williamson, W.M., Walshe, G., Hunter, K.A., 2009. Modifying the Surface Charge of Pathogen-Sized Microspheres for Studying Pathogen Transport in Groundwater. *J. Environ. Qual.* 38 (6), 2210–2217.
- Paul, D., Kolar, P., Hall, S.G., 2021. A review of the impact of environmental factors on the fate and transport of coronaviruses in aqueous environments. *npj Clean. Water* 4 (1), 7.
- Pelley, A.J., Tufenki, N., 2008. Effect of particle size and natural organic matter on the migration of nano- and microscale latex particles in saturated porous media. *J. Colloid Interface Sci.* 321 (1), 74–83.
- Pennington, H., 2010. *Escherichia coli* O157. *Lancet* 376 (9750), 1428–1435.
- Penrod, S.L., Olson, T.M., Grant, S.B., 1996. Deposition Kinetics of Two Viruses in Packed Beds of Quartz Granular Media. *Langmuir* 12 (23), 5576–5587.
- Pinon, A., Vialette, M., 2018. Survival of Viruses in Water. *Intervirology* 61 (5), 214–222.
- Predeus, D., Lassabatere, L., Louis, C., Gehan, H., Brichart, T., Winiarski, T., Angulo-Jaramillo, R., 2017. Nanoparticle transport in water-unsaturated porous media: effects of solution ionic strength and flow rate. *J. Nanopart. Res.* 19 (3), 104.
- Qin, Y.Q., 2020. Mechanism of virus-colloid cotransport in saturated/unsaturated porous media. Master Thesis, Jilin University. (in Chinese with English abstract).
- Qin, Y.Q., Wen, Z., Zhang, W.J., Chai, J.F., Liu, D., Wu, S.Y., 2020. Different roles of silica nanoparticles played in virus transport in saturated and unsaturated porous media. *Environ. Pollut.* 259, 113861.
- Randazzo, W., Truchado, P., Cuevas-Ferrando, E., Simon, P., Allende, A., Sanchez, G., 2020. SARS-CoV-2 RNA in wastewater anticipated COVID-19 occurrence in a low prevalence area. *Water Res.* 181, 115942.
- Rashid, M., Pandit, D., 2017. Determination of appropriate service quality attributes for household toilets in rural settlements of India based on user perception. *Environ. Dev. Sustainability* 19 (4), 1381–1406.
- Rasmuson, A., VanNess, K., Ron, C.A., Johnson, W.P., 2019. Hydrodynamic versus Surface Interaction Impacts of Roughness in Closing the Gap between Favorable and Unfavorable Colloid Transport Conditions. *Environ. Sci. Technol.* 53 (5), 2450–2459.
- Rawlinson, L.-A.-B., O'Gara, J.P., Jones, D.S., Brayden, D.J., 2011. Resistance of *Staphylococcus aureus* to the cationic antimicrobial agent poly(2-(dimethylamino ethyl)methacrylate) (pDMAEMA) is influenced by cell-surface charge and hydrophobicity. *J. Med. Microbiol.* 60 (7), 968–976.
- Redman, J., A., Grant, and, S., B., 1997. Filtration of recombinant Norwalk virus particles and bacteriophage MS2 in quartz sand. *Environ. Sci. Technol.* 31 (12), 3378–3383.
- Ren, L.Y., Hong, Z.N., Liu, Z.D., Xu, R.K., 2018. ATR-FTIR investigation of mechanisms of *Bacillus subtilis* adhesion onto variable- and constant-charge soil colloids. *Colloid Surface B* 162, 288–295.
- Rhoden, K., Alonso, J., Carmona, M., Pham, M., Barnes, A.N., 2021. Twenty years of waterborne and related disease reports in Florida, USA. *One Health* 13, 100294.
- Rizzo, L., Manaia, C., Merlin, C., Schwartz, T., Dagot, C., Ploy, M.C., Michael, I., Fatta-Kassinos, D., 2013. Urban wastewater treatment plants as hotspots for antibiotic resistant bacteria and genes spread into the environment: A review. *Sci. Total Environ.* 447, 345–360.
- Robertson, J.B., Edberg, S.C., 1997. Natural Protection of Spring and Well Drinking Water Against Surface Microbial Contamination. I. Hydrogeological Parameters. *Crit. Rev. Microbiol.* 23 (2), 143–178.
- Rodrigues, L.R., Banat, I.M., Mei, H.C., Teixeira, J.A., Oliveira, R., 2006. Interference in adhesion of bacteria and yeasts isolated from explanted voice prostheses to silicone rubber by rhamnolipid biosurfactants. *J. Appl. Microbiol.* 100 (3), 470–480.
- Rong, X.M., Huang, Q.Y., Chen, W.L., 2007. Microcalorimetric investigation on the metabolic activity of *Bacillus thuringiensis* as influenced by kaolinite, montmorillonite and goethite. *Appl. Clay Sci.* 38 (1–2), 97–103.
- Rong, X.M., 2008. Thermodynamic investigations on the interactions of bacteria with soil clay minerals, PhD Thesis, Huazhong Agricultural University (in Chinese with English abstract)
- Rutjes, S.A., Italiaander, R., van den Berg, H., Lodder, W.J., Husman, A.M.D., 2005. Isolation and detection of enterovirus RNA from large-volume water samples by using the NucliSens miniMAG system and real-time nucleic acid sequence-based amplification. *Appl. Environ. Microbiol.* 71 (7), 3734–3740.
- Signoretto, C., del Mar Lleo, M., Tafi, M.C., Canepari, P., 2000. Cell Wall Chemical Composition of *Enterococcus faecalis* in the Viable but Nonculturable State. *Appl. Environ. Microbiol.* 66, 1953–1959.
- Salerno, M.B., Flamm, M., Logan, B.E., Velegol, D., 2006. Transport of rodlike colloids through packed beds. *Environ. Sci. Technol.* 40 (20), 6336–6340.
- Samandoulgou, I., Hammami, R., Rayas, R.M., Fliss, I., Jean, J., 2015. Stability of Secondary and Tertiary Structures of Virus-Like Particles Representing Noroviruses: Effects of pH, Ionic Strength, and Temperature and Implications for Adhesion to Surfaces. *Appl. Environ. Microbiol.* 81 (22), 7680–7686.
- Sarkar, M., Saha, S., 2020. Structural insight into the role of novel SARS-CoV-2 E protein: A potential target for vaccine development and other therapeutic strategies. *PLoS ONE* 15, (8) e0237300.
- Sasidharan, S., Bradford, S.A., Torkzaban, S., Ye, X., Vanderzalm, J., Du, X., Page, D., 2017a. Unraveling the complexities of the velocity dependency of *E. coli* retention and release parameters in saturated porous media. *Sci. Total Environ.* 603–604, 406–415.
- Sasidharan, S., Torkzaban, S., Bradford, S.A., Cook, P.G., Gupta, V.V.S.R., 2017b. Temperature dependency of virus and nanoparticle transport and retention in saturated porous media. *J. Contam. Hydrol.* 196, 10–20.
- Sasidharan, S., Torkzaban, S., Bradford, S.A., Kookana, R., Page, D., Cook, P.G., 2016. Transport and retention of bacteria and viruses in biochar-amended sand. *Sci. Total Environ.* 548–549, 100–109.
- Scheller, C., Krebs, F., Minkner, R., Astner, I., Gil-Moles, M., Wätzig, H., 2020. Physicochemical properties of SARS-CoV-2 for drug targeting, virus inactivation

- and attenuation, vaccine formulation and quality control. *Electrophoresis* 41 (13–14), 1137–1151.
- Schijven, J.F., Hassanzadeh, S.M., 2000. Removal of Viruses by Soil Passage: Overview of Modeling, Processes, and Parameters. *Crit. Rev. Env. Sci. Technol.* 30 (1), 49–127.
- Schinner, T., Letzner, A., Liedtke, S., Castro, F.D., Eydelenant, I.A., Tufenkji, N., 2010. Transport of selected bacterial pathogens in agricultural soil and quartz sand. *Water Res.* 44 (4), 1182–1192.
- Schulze-Makuch, D., Guan, H.D., Pillai, S.D., 2003. Effects of pH and geological medium on bacteriophage MS2 transport in a model aquifer. *Geomicrobiol. J.* 20 (1), 73–84.
- Seymour, M.B., Chen, G.X., Su, C.M., Li, Y.S., 2013. Transport and Retention of Colloids in Porous Media: Does Shape Really Matter? *Environ. Sci. Technol.* 47 (15), 8391–8398.
- Shellenberger, K., Logan, B.E., 2002. Effect of Molecular Scale Roughness of Glass Beads on Colloidal and Bacterial Deposition. *Environ. Sci. Technol.* 36 (2), 184–189.
- Shen, C.Y., Jin, Y., Zhuang, J., Li, T.T., Xing, B.S., 2020. Role and importance of surface heterogeneities in transport of particles in saturated porous media. *Crit. Rev. Env. Sci. Technol.* 50 (3), 244–329.
- Shen, C.Y., Bradford, S.A., 2021. Why Are Viruses Spiked? *Mosphere* 6 (1), e01339–01320.
- Shen, C.Y., Bradford, S.A., Li, T.T., Li, B.G., Huang, Y.F., 2018. Can nanoscale surface charge heterogeneity really explain colloid detachment from primary minima upon reduction of solution ionic strength? *J. Nanopart. Res.* 20 (6), 1–18.
- Shen, C.Y., Wang, F., Li, B.G., Jin, Y., Wang, L.P., Huang, Y.F., 2012. Application of DLVO Energy Map To Evaluate Interactions between Spherical Colloids and Rough Surfaces. *Langmuir* 28 (41), 14681–14692.
- Shen, C.Y., Wu, L., Zhang, S.W., Ye, H.C., Li, B.G., Huang, Y.F., 2014. Heteroaggregation of microparticles with nanoparticles changes the chemical reversibility of the microparticles' attachment to planar surfaces. *J. Colloid Interface Sci.* 421, 103–113.
- Sherchan, S.P., Shahin, S., Ward, L.M., Tandukar, S., Kitajima, M., 2020. First detection of SARS-CoV-2 RNA in wastewater in North America: A study in Louisiana, USA. *Sci. Total Environ.* 743, 140621.
- Shields, P.A., Farrah, S.R., 1983. Influence of salts on electrostatic interactions between poliovirus and membrane filters. *Appl. Environ. Microbiol.* 45 (2), 526–531.
- Shields, P.A., 1986. Factors influencing virus adsorption to solids. University of Florida, Gainesville. Ph.D. thesis.
- Shin, H., Park, H., Seo, D.J., Jung, S., Yeo, D., Wang, Z., Park, K.H., Choi, C., 2019. Foodborne Viruses Detected Sporadically in the Fresh Produce and Its Production Environment in South Korea. *Foodborne Pathog. Dis.* 16 (6), 411–420.
- Sinton, L.W., Mackenzie, M.L., Karki, N., Braithwaite, R.R., Hall, C.H., Flintoft, M.J., 2010. Transport of Escherichia coli and F-RNA bacteriophages in a 5 m column of saturated pea gravel. *J. Contam. Hydrol.* 117 (1), 71–81.
- Sirivithayapakorn, S., Keller, A., 2003. Transport of colloids in unsaturated porous media: A pore-scale observation of processes during the dissolution of air-water interface. *Water Resour. Res.* 39 (12), 1346.
- Sosnowski, T.R., 2021. Inhaled aerosols: their role in COVID-19 transmission including biophysical interactions in the lungs. *Curr. Opin. Colloid Interface Sci.*, 101451.
- Sun, J., Zhu, A.R., Li, H.Y., Zheng, K., Zhuang, Z., Chen, Z., Shi, Y.X., Zhang, Z.Y., Chen, S.-B., Liu, X.S., Dai, J., Li, X.B., Huang, S.X., Huang, X.F., Luo, L., Wen, L.Y., Zhuo, J.F., Li, Y.M., Wang, Y.Q., Zhang, L., Zhang, Y.J., Li, F., Feng, L.Q., Chen, X.W., Zhong, N., S., Yang, Z.F., Huang, J.C., Zhao, J.C., Li, Y.-M., 2020. Isolation of infectious SARS-CoV-2 from urine of a COVID-19 patient. *Emerging Microbes Infect.* 9 (1), 991–993.
- Syngouva, V.I., Chrysikopoulos, C.V., 2016. Cotransport of clay colloids and viruses through water-saturated vertically oriented columns packed with glass beads: Gravity effects. *Sci. Total Environ.* 545, 210–218.
- Syngouva, V.I., Chrysikopoulos, C.V., Kokkinos, P., Tselepi, M.A., Vantarakis, A., 2017. Cotransport of human adenoviruses with clay colloids and TiO<sub>2</sub> nanoparticles in saturated porous media: Effect of flow velocity. *Sci. Total Environ.* 598, 160–167.
- Tong, M.P., Ding, J.L., Shen, Y., Zhu, P.T., 2010a. Influence of biofilm on the transport of fullerene (C60) nanoparticles in porous media. *Water Res.* 44 (4), 1094–1103.
- Tong, M.P., Long, G.Y., Jiang, X.J., Kim, H.N., 2010b. Contribution of Extracellular Polymeric Substances on Representative Gram Negative and Gram Positive Bacterial Deposition in Porous Media. *Environ. Sci. Technol.* 44 (7), 2393–2399.
- Torkzaban, S., Bradford, S.A., 2016. Critical role of surface roughness on colloid retention and release in porous media. *Water Res.* 88, 274–284.
- Torkzaban, S., Bradford, S.A., Vanderzalm, J.L., Patterson, B.M., Harris, B., Prommer, H., 2015. Colloid release and clogging in porous media: Effects of solution ionic strength and flow velocity. *J. Contam. Hydrol.* 181, 161–171.
- Torkzaban, S., Hassanzadeh, S., Schijven, J., De Bruin, H., de Roda Husman, A., 2006. Virus transport in saturated and unsaturated sand columns. *Vadose Zone J.* 5 (3), 877–885.
- Torkzaban, S., Hocking, M., Bradford, S.A., Tazehkand, S.S., Sasidharan, S., Šimunek, J., 2019. Modeling Virus Transport and Removal during Storage and Recovery in Heterogeneous Aquifers. *J. Hydrol.* 578, 124082.
- Tsuneda, S., Aikawa, H., Hayashi, H., Yuasa, A., Hirata, A., 2003. Extracellular polymeric substances responsible for bacterial adhesion onto solid surface. *FEMS Microbiol. Lett.* 223 (2), 287–292.
- Tufenkji, N., 2007. Colloid and Microbe Migration in Granular Environments: A Discussion of Modelling Methods, In *Colloidal transport in porous media. Colloidal Transport in Porous Media*, 119–142.
- Van Loosdrecht, M.C., Lyklema, J., Norde, W., Schraa, G., Zehnder, A.J., 1987. Electrophoretic mobility and hydrophobicity as a measured to predict the initial steps of bacterial adhesion. *Appl. Environ. Microbiol.* 53 (8), 1898–1901.
- Vanderzalm, J.L., Page, D.W., Barry, K.E., Dillon, P.J., 2010. A comparison of the geochemical response to different managed aquifer recharge operations for injection of urban stormwater in a carbonate aquifer. *Appl. Geochem.* 25 (9), 1350–1360.
- Walker, S.L., Redman, J.A., Elimelech, M., 2004. Role of Cell Surface Lipopolysaccharides in *Escherichia coli*K12 Adhesion and Transport. *Langmuir* 20 (18), 7736–7746.
- Walshe, G.E., Pang, L., Flury, M., Close, M.E., Flintoft, M., 2010. Effects of pH, ionic strength, dissolved organic matter, and flow rate on the co-transport of MS2 bacteriophages with kaolinite in gravel aquifer media. *Water Res.* 44 (4), 1255–1269.
- Wan, J., Tokunaga, T.K., 2002. Partitioning of clay colloids at air–water interfaces. *J. Colloid Interface Sci.* 247 (1), 54–61.
- Wang, L.X., Xu, S.P., Li, J., 2011. Effects of Phosphate on the Transport of *Escherichia coli* O157:H7 in Saturated Quartz Sand. *Environ. Sci. Technol.* 45 (22), 9566–9573.
- Wang, Y.Y., Hammes, F., Duggelin, M., Egli, T., 2008. Influence of size, shape, and flexibility on bacterial passage through micropore membrane filters. *Environ. Sci. Technol.* 42 (17), 6749–6754.
- Wardlaw, T., Salama, P., Brocklehurst, C., Chopra, M., Mason, E., 2010. Diarrhoea: why children are still dying and what can be done. *Lancet* 375 (9718), 870–872.
- Weidhaas, J., Garner, E., Basden, T., Harwood, V.J., 2014. Run-off studies demonstrate parallel transport behaviour for a marker of poultry fecal contamination and *Staphylococcus aureus*. *J. Appl. Microbiol.* 117 (2), 417–429.
- Weiss, T.H., Mills, A.L., Hornberger, G.M., Herman, J.S., 1995. Effect of Bacterial Cell Shape on Transport of Bacteria in Porous Media. *Environ. Sci. Technol.* 29 (7), 1737–1740.
- Weststrate, J., Dijkstra, G., Eshuis, J., Gianoli, A., Rusca, M., 2019. The Sustainable Development Goal on Water and Sanitation: Learning from the Millennium Development Goals. *Soc. Indic. Res.* 143 (2), 795–810.
- White, K., Dickson-Anderson, S., Majury, A., McDermott, K., Hynds, P., Brown, R.S., Schuster-Wallace, C., 2021. Exploration of *E. coli* contamination drivers in private drinking water wells: An application of machine learning to a large, multivariable, geo-spatio-temporal dataset. *Water Res.* 197, 117089.
- Whitehead, K., McCue, K.A., 2010. Virucidal efficacy of disinfectant actives against feline calicivirus, a surrogate for norovirus, in a short contact time. *Am. J. Infect. Control* 38 (1), 26–30.
- Widdowson, M.A., Monroe, S.S., Glass, R.I., 2005. Are noroviruses emerging? *Emerg. Infect. Dis.* 11 (5), 735–737.
- Wolfe, M.K., Gallandat, K., Daniels, K., Desmarais, A.M., Scheinman, P., Lantagne, D., 2017. Handwashing and Ebola virus disease outbreaks: A randomized comparison of soap, hand sanitizer, and 0.05% chlorine solutions on the inactivation and removal of model organisms Phi6 and *E. coli* from hands and persistence in rinse water. *PLoS ONE* 12, (2) e0172734.
- Wong, K., Bouchard, D., Molina, M., 2014. Relative transport of human adenovirus and MS2 in porous media. *Colloids Surf. B* 122, 778–784.
- Wu, C.R., Liu, Y., Yang, Y.Y., Zhang, P., Zhong, W., Wang, Y.L., Wang, Q.Q., Xu, Y., Li, M., X., Li, X.Z., Zheng, M.Z., Chen, L.X., Li, H., 2020. Analysis of therapeutic targets for SARS-CoV-2 and discovery of potential drugs by computational methods. *Acta Pharm. Sin. B* 10 (5), 766–788.
- Wu, G., Yang, J., Jiang, H.C., Deng, Y.M., Lear, G., 2019. Distribution of potentially pathogenic bacteria in the groundwater of the Jianghan Plain, central China. *Int. Biodeterior. Biodegrad.* 143, 104711.
- Wu, S., 2021. Study on environmental behavior characteristics of *Escherichia coli* O157:H7 in the porous media in the presence of composite colloid. Jilin University (in Chinese with English abstract).
- Xu, L.-C., Logan, B.E., 2006. Adhesion forces between functionalized latex microspheres and protein-coated surfaces evaluated using colloid probe atomic force microscopy. *Colloids Surf. B* 48 (1), 84–94.
- Xu, L.-C., Vadillo-Rodriguez, V., Logan, B.E., 2005. Residence Time, Loading Force, pH, and Ionic Strength Affect Adhesion Forces between Colloids and Biopolymer-Coated Surfaces. *Langmuir* 21 (16), 7491–7500.
- Yan, L.J., Herrmann, M., Kampe, B., Lehmann, R., Totsche, K.U., Kusel, K., 2020. Environmental selection shapes the formation of near-surface groundwater microbiomes. *Water Res.* 170, 115341.
- Yang, H.Y., Ge, Z., Wu, D., Tong, M.P., Ni, J.R., 2016. Cotransport of bacteria with hematite in porous media: Effects of ion valence and humic acid. *Water Res.* 88, 586–594.
- Yang, H.Y., Kim, H., Tong, M.P., 2012a. Influence of humic acid on the transport behavior of bacteria in quartz sand. *Colloids Surf. B* 91, 122–129.
- Yang, H.Y., Tong, M.P., Kim, H., 2012b. Influence of Bentonite Particles on Representative Gram Negative and Gram Positive Bacterial Deposition in Porous Media. *Environ. Sci. Technol.* 46 (21), 11627–11634.
- Yates, M.V., Gerba, C.P., Kelley, L.M., 1985. Virus persistence in groundwater. *Appl. Environ. Microbiol.* 49 (4), 778–781.
- Ye, Y., Ellenberg, R.M., Graham, K.E., Wigginton, K.R., 2016. Survivability, Partitioning, and Recovery of Enveloped Viruses in Untreated Municipal Wastewater. *Environ. Sci. Technol.* 50 (10), 5077–5085.

- Yu, X.P., 2016. Study on Viruses Transportation though Saturated Porous Media under Managed Groundwater Recharge Condition. Master Thesis, Jilin University (in Chinese with English abstract).
- Zeraik, A.E., Nitschke, M., 2010. Biosurfactants as agents to reduce adhesion of pathogenic bacteria to polystyrene surfaces: effect of temperature and hydrophobicity. *Curr. Microbiol.* 61 (6), 554–559.
- Zhang, D.Y., Ling, H.B., Huang, X., Li, J., Li, W.W., Yi, C., Zhang, T., Jiang, Y.Z., He, Y.N., Deng, S.Q., Zhang, X., Wang, X.Z., Liu, Y., Li, G.H., Qu, J.H., 2020. Potential spreading risks and disinfection challenges of medical wastewater by the presence of Severe Acute Respiratory Syndrome Coronavirus 2 (SARS-CoV-2) viral RNA in septic tanks of Fangcang Hospital. *Sci. Total Environ.* 741, 140445.
- Zhang, D.Y., Yang, Y.F., Li, M., Lu, Y., Liu, Y., Jiang, J.K., Liu, R.P., Liu, J.G., Huang, X., Li, G.H., Qu, J.H., 2021a. Ecological Barrier Deterioration Driven by Human Activities Poses Fatal Threats to Public Health due to Emerging Infectious Diseases. *Eng.*, DOI10.1016/j.eng.2020.1011.1002.
- Zhang, D.Y., Zhang, X., Yang, Y.F., Huang, X., Jiang, J.K., Li, M., Ling, H.B., Li, J., Liu, Y., Li, G.H., Li, W.W., Yi, C., Zhang, T., Jiang, Y.Z., Xiong, Y., Hu, Z.Y., Wang, X.Z., Deng, S.Q., Zhao, P., Qu, J.H., 2021b. SARS-CoV-2 spillover into hospital outdoor environments. *J. Hazard. Mater. Lett.* 2, 100027.
- Zhang, D.Y., Zhang, X.L., Ma, R., Deng, S.Q., Wang, X.Z., Wang, X.Q., Zhang, X., Huang, X., Liu, Y., Li, G.H., Qu, J.H., Zhu, Y., Li, J.Y., 2021c. Ultra-fast and onsite interrogation of Severe Acute Respiratory Syndrome Coronavirus 2 (SARS-CoV-2) in waters via surface enhanced Raman scattering (SERS). *Water Res.* 200, 117243.
- Zhang, H., Zhang, J., Zhao, B., Zhang, C., 2010. Removal of bacteriophages MS2 and phiX174 from aqueous solutions using a red soil. *J. Hazard. Mater.* 180 (1–3), 640–647.
- Zhang, M., Ghosh, S., Kumar, M., Santiana, M., Bleck, C.K.E., Chaimongkol, N., Altan-Bonnet, N., Shuai, D., 2021d. Emerging Pathogenic Unit of Vesicle-Cloaked Murine Norovirus Clusters is Resistant to Environmental Stresses and UV254 Disinfection. *Environ. Sci. Technol.*, 6197–6205.
- Zhu, N., Zhang, D., Wang, W., Li, X., Yang, B., Song, J., Zhao, X., Huang, B., Shi, W., Lu, R., Niu, P., Zhan, F., Ma, X., Wang, D., Xu, W., Wu, G., Gao, G.F., Tan, W., 2020. A Novel Coronavirus from Patients with Pneumonia in China, 2019. *New Engl. J. Med.* 382 (8), 727–733.
- Zhang, W.J., Wu, S.Y., Qin, Y.Q., Li, S., Lei, L.C., Sun, S.M., Yang, Y.S., 2021e. Deposition and mobilization of viruses in unsaturated porous media: Roles of different interfaces and straining. *Environ. Pollut.* 270, 116072.
- Zhang, W.J., Li, S., Wang, S., Lei, L.C., Yu, X.P., Ma, T.Y., 2018. Transport of Escherichia coli phage through saturated porous media considering managed aquifer recharge. *Environ. Sci. Pollut. Res.* 25 (7), 6497–6513.
- Zhao, W.Q., Walker, S.L., Huang, Q.Y., Cai, P., 2014. Adhesion of bacterial pathogens to soil colloidal particles: Influences of cell type, natural organic matter, and solution chemistry. *Water Res.* 53, 35–46.
- Zhong, H., Jiang, Y.B., Zeng, G.M., Liu, Z.F., Liu, L.X., Liu, Y., Yang, X., Lai, M.Y., He, Y.B., 2015. Effect of low-concentration rhamnolipid on adsorption of *Pseudomonas aeruginosa* ATCC 9027 on hydrophilic and hydrophobic surfaces. *J. Hazard. Mater.* 285, 383–388.
- Zhuang, J., Jin, Y., 2003a. Virus retention and transport as influenced by different forms of soil organic matter. *J. Environ. Qual.* 32 (3), 816–823.
- Zhuang, J., Jin, Y., 2003b. Virus retention and transport through Al-oxide coated sand columns: effects of ionic strength and composition. *J. Contam. Hydrol.* 60 (3–4), 193–209.
- Zimmerman, A.E., Howard-Varona, C., Needham, D.M., John, S.G., Worden, A.Z., Sullivan, M.B., Waldbauer, J.R., Coleman, M.L., 2020. Metabolic and biogeochemical consequences of viral infection in aquatic ecosystems. *Nat. Rev. Microbiol.* 18 (1), 21–34.
- Zmantar, T., Bettaieb, F., Chaieb, K., Ezzili, B., Mora-Ponsonnet, L., Othmane, A., Jaffrezic, N., Bakhrout, A., 2011. Atomic force microscopy and hydrodynamic characterization of the adhesion of *staphylococcus aureus* to hydrophilic and hydrophobic substrata at different pH values. *World J. Microbiol. Biotechnol.* 27 (4), 887–896.